REPORT



WASTE MANAGEMENT OF CANADA CORPORATION

WATFORD, ONTARIO

TWIN CREEKS ENVIRONMENTAL CENTRE: 2022 FOURTH QUARTER & ANNUAL MONITORING REPORT VOLUME 1 OF 5 – COMPLIANCE MONITORING

RWDI #2202861.01 March 1, 2023

SUBMITTED TO

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March 1, 2023

Ms. Angela McLachlan Environmental Compliance Manager Waste Management of Canada Corporation Twin Creeks Environmental Centre 5678 Nauvoo Road (Watford) Warwick Township, County of Lambton

Re: Waste Management of Canada Corporation
2022 Fourth Quarter and Annual Monitoring Report
Twin Creeks Environmental Centre, Warwick Township, County of Lambton, Ontario
Volumes 1 and 2 - Text, Figures, Tables, and Appendices

Dear Ms. McLachlan.

RWDI AIR Inc. (RWDI) is pleased to provide the 2022 Fourth Quarter and Annual Monitoring Report, **Volumes 1** and **2:** Compliance Monitoring and Operations Program, for the Twin Creeks Environmental Centre. **Volume 1** includes the text, figures, and tables. **Volume 2** includes the appendices, which contain the technical information and supporting documentation for the Compliance Monitoring and Operations Program and is provided in two (2) bound sections. **Volume 3** – Poplar System Monitoring Program, **Volume 4** – Ambient Air Quality Monitoring Program, and **Volume 5** – Noise Monitoring Program, are provided under separate covers

In November 2010, the Ministry of the Environment (MOE) issued the Technical Guidance Document entitled "Monitoring and Reporting for Waste Disposal Sites, Groundwater and Surface Water" (MOE, 2010). Appended to this report is a completed Monitoring and Screening Checklist from the above Technical Guidance Document, which provides certification of the Competent Environmental Practitioner (CEP). The Monitoring and Screening Checklist is provided in **Appendix S**.

Volumes 1 and 2 of the 2022 Fourth Quarter and Annual Monitoring Report have been prepared in consideration of Conditions 15.4 through 15.7 of amended Environmental Compliance Approval No. A032203, dated December 19, 2020 (Waste ECA), and provides a detailed interpretive analysis of the 2022 findings for the compliance monitoring at the Twin Creeks Environmental Centre and a summary of its operations in 2022.

We trust that this report satisfies your requirements. Please contact us if you have any questions.

Yours truly,

RWDI AIR Inc.

Brent J. Langille, B.Sc., P.Geo. Technical Director | Principal bjl/kta

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### 1 INTRODUCTION

### 1.1 Purpose & Scope

The purpose of the 2022 Compliance Monitoring and Operations Program for Waste Management of Canada Corporation's (WM) Twin Creeks Environmental Centre (Site) is as follows.

- To report compliance with the terms and conditions of the relevant Environmental Compliance Approvals (ECAs) (waste, sewage, and air) and regulatory permits (PTTW) for the Site, the landfill gas flare systems, and stormwater management facilities for the reporting period of January 1 to December 31, 2022.
- To satisfy Conditions 5, 6, and 7 of the Environmental Assessment Act approval, dated January 15, 2007.
- To assess potential effects of the landfill on groundwater and surface water quality.
- To evaluate the monitoring program(s) on an annual basis and to recommend improvements as operations at the Site mature.
- To determine the need for the implementation of contingency measures.

**Volumes 1 and 2** of the 2022 Fourth Quarter and Annual Monitoring Report (2022 Annual Report) satisfy the reporting requirements for both the fourth quarter (Q4) of 2022 (October 1 to December 31) and the 2022 compliance monitoring and operations calendar year monitoring period. Reporting involves a data collection component and a comprehensive analysis and interpretation component. The 2022 data were collected by WM and RWDI AIR Inc. (RWDI), while the data compiled for **Volumes 1 and 2** of the 2022 Annual Report were collated and analyzed by RWDI. The following presents a summary of the data collection responsibilities for the Compliance Monitoring Program in 2022.

#### **Waste Management of Canada Corporation:**

- Operations Logs (Dust and Litter)
- Leachate Volume Tracking
- Tonnage Tracking Information (Waste & Recyclables)
- Water Taking Information
- Complaint Logs
- Collection of Daily Liquid Levels from Expansion Site PDL Pumping Stations

#### **RWDI AIR Inc.:**

- Liquid Level Monitoring (Leachate and Groundwater)
- Groundwater Quality Monitoring
- Surface Water Quality Monitoring
- Contaminated Soil & ASR Quality Monitoring
- Leachate Quality Monitoring
- Monthly Collection of Liquid levels from Expansion Site SDL Pumping Stations
- Monthly Site Inspections
- · Overall Data Collation and Reporting

Air quality, noise, total suspended particulate and Poplar System monitoring, and reporting were also completed by RWDI in 2022. The monitoring results, analysis, and interpretation for the Poplar System Monitoring Program are presented in **Volume 3** of the 2022 Annual Report. Approval to discontinue the monitoring of the Poplar Plantation was received on February 20, 2013, per Notice No. 1 of the Sewage ECA. Therefore, a relevant volume is not included in this Annual Report.

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### 1.2 Site Regulatory Framework

The Site is owned and is operated by WM in conformance with the regulatory approvals noted below. Copies of the relevant ECAs and their amendments, as well as the PTTW, are provided in **Appendix A**.

- Amended Environmental Compliance Approval (ECA) No. A032203, dated February 4, 2023. It is noted
  that during the 2022 calendar year, WM was required to conform to the ECA dated December 19, 2020
  (Waste ECA).
- Amended ECA for Industrial Sewage Works No. 2403-BE6LZ4, dated August 21, 2019 (Sewage ECA).
- Amended ECA for Air No. 4155-BMCLZ8, dated March 3, 2020 (Air ECA).
- Permit-To-Take-Water (PTTW) No. 4682-BLJRYJ, dated November 8, 2021, for the removal of surface water from four (4) Sedimentation Ponds and the dewatering of the Secondary Drainage Layer (SDL) for the Expansion Site.
- MECP Letter entitled "Request for Modification to Surface Water Monitoring/Assessment Process at Twin Creeks Landfill", dated February 27, 2014 (2014 MECP Letter).

Throughout 2022, quarterly monitoring reports were submitted to relevant stakeholders in accordance with Condition 15.4 of the Waste ECA. **Volumes 1 and 2** of the 2022 Annual Report have been prepared in consideration of Conditions 15.4 through 15.7 of the Waste ECA and Condition 13(4) of the Sewage ECA. **Volumes 1 and 2** provide a detailed analysis of the 2022 findings for the compliance monitoring at the Site and its operation in 2022.

This 2022 Annual Report is also prepared in accordance with Conditions 5, 6, and 7 of the Notice of Approval to Proceed with the Undertaking, dated January 15, 2007, in regard to the Environmental Assessment Act (EAA) approval of the Warwick Landfill Expansion (renamed Twin Creeks Landfill and subsequently to Twin Creeks Environmental Centre).

### 1.3 Site Description & Background

The Site is a solid, non-hazardous waste landfill site that contains an old landfill (Existing Site) and an area approved for expansion (Expansion Site). The Site is located on Part of Lots 19 and 20, Concession 3, south of Egremont Road (SER), and part of Lots 20 to 22, Concession 4 SER, in the Township of Warwick, County of Lambton, Ontario. The Site occupies an area of 301 hectares (ha) with 101.8 ha permitted for landfilling. A Site Location Map is presented on **Figure 1**.

Landfill waste at the Site has two (2) distinct waste disposal areas: 1) the Existing Site; and 2) the Expansion Site. The Existing Site is divided into waste cells as presented on **Figure 2**. The progression of construction of the Expansion Site is summarized below and presented on **Figure 2**.



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Expansion Site Cell	Date of First Waste Tonnage Deposited
Cell 1A Stage I	November 16, 2009
Cell 1A Stage II	September 21, 2010
Cell 1B Stage I	August 18, 2011
Cell 1B Stage II	June 20, 2012
Cell 2A	November 18, 2013
Cell 2B	September 25, 2014
Cell 2C	September 10, 2015
Cell 2D	August 19, 2016
Cell 2E	September 8, 2017
Cell 4A Stage 1	September 25, 2019
Cell 4A Stage 2	November 14, 2019
Cell 4B Stage 1	September 22, 2020
Cell 4B Stage 2 November 23, 2020	
Cell 4C	August 10, 2021
Cell 6A	September 14, 2022

During 2022, waste disposal occurred in Cell 4 and Cell 6A of the Expansion Site. Waste disposal in Cell 6A commenced on September 14, 2022, as the landfill liner system was approved to accept waste. Cell 6B pre-excavation activities related to the future construction of the Cell 6A landfill liner system of the Expansion Site were completed in 2022. Upon completion of the landfill liner system of Cell 6B, waste disposal is scheduled to also occur Cell 6B in 2023.

The South Cell (formerly South Fill Area) of the Existing Site consists of historical waste landfilled into trenches of various widths and depths excavated into the clayey soil. There are three (3) finger drains that transect through the trenches to provide gravity drainage toward the perimeter maintenance holes (MH16, MH17, and MH18) for leachate collection. The remainder of the waste footprint of the Existing Site also contains historical waste trenches but includes newer waste cells constructed with re-compacted clayey liners and, in some cells, waste underdrains.

Waste within the Expansion Site will extend to an average design depth of 15 m below existing grade. Leachate is managed with a liner system that directs leachate toward dedicated leachate pumping stations. The liner system consists of a primary drainage layer (PDL) below the waste to convey leachate, an underlying re-compacted clayey liner (primary liner), then a secondary drainage layer (SDL) for leak detection and contingency use, which is all underlain by a thick natural barrier of clayey soil. Design specifications are provided under Items 66 to 68 in Schedule A of the Waste ECA, namely, the report titled "Development and Operations Plans – Warwick Landfill Expansion (Volumes 1 to 3)", as prepared by Henderson Paddon & Associates and dated March 2008 (D&O Report).

Leachate generated at the Site can be managed either on-Site through irrigation for poplar trees to be treated by phytoremediation or by transport for off-Site treatment and disposal. The Poplar System was decommissioned in June 2014 as part of the construction activities to expand the footprint from the previous 3.3 ha to 9 ha. Leachate phytoremediation is approved for the expanded Poplar System located within the waste footprint of the Existing

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Site, as shown on **Figure 2**. In 2022, irrigation liquid was applied to the poplar trees intermittently from May 11 to October 11. It should be noted that the Poplar System pertains to a plot of poplar trees planted on the landfill cap of the Existing Site, whereas the Poplar Plantation refers to the plot of trees planted on native soil and is located south of the Existing Site. Details pertaining to the Poplar System Monitoring Program completed in 2022, including surface water monitoring in response to precipitation events of ≥ 10 millimetres (mm) in 24 hours, are provided in **Volume 3** of the 2022 Annual Report.

Surface water flow is ephemeral, with flow that typically occurs after snowmelt or prolonged periods of precipitation. Details on the surface water quality system are provided in **Section 2.4**. In summary, surface water is managed through a network of on-Site ditches, a Temporary Water Storage Area, and four (4) on-Site Sedimentation Ponds. Surface water runoff from the Site flows toward: 1) Kersey Drain (Brown Creek) to the east, and 2) to drains and ditches associated with Bear Creek to the west.

Assessments of local groundwater and surface water have continued to achieve acceptable quality at the Site compliance boundaries. A list of historical monitoring reports is provided in **Table B-1**, **Appendix B**.

For reference to the Site construction design details, **Appendix L** contains construction details of the Existing and Expansion Sites, with the breakdown as outlined below.

#### **Existing Site:**

The construction details: "Laidlaw Waste Systems (Warwick) Limited, Warwick Township Landfill Site, Development and Operations Plan" from the Development & Operations Report, Canadian Waste Services Inc., Warwick Landfill, Warwick Township (Henderson, Paddon Environmental Inc., October 1997), which is Item 37 of Schedule A of the Waste ECA (included as **Appendix L-1**, **Appendix L**).

#### **Expansion Site:**

The construction details: "Drawings 111 – 120, 125, and 127" from the Development & Operations Plans, Warwick Landfill Expansion, Volume 1 of 3, which is Item 66 of Schedule A of the Waste ECA (included as **Appendix L-2, Appendix L**).

### As-built Information for Waste Cells Constructed in 2022:

The CQA/CQC Liner System Summary Report (Cell 6A), as prepared by RWDI, dated September 12, 2022, respectively, can be found in **Appendix L-3**, **Appendix L**.

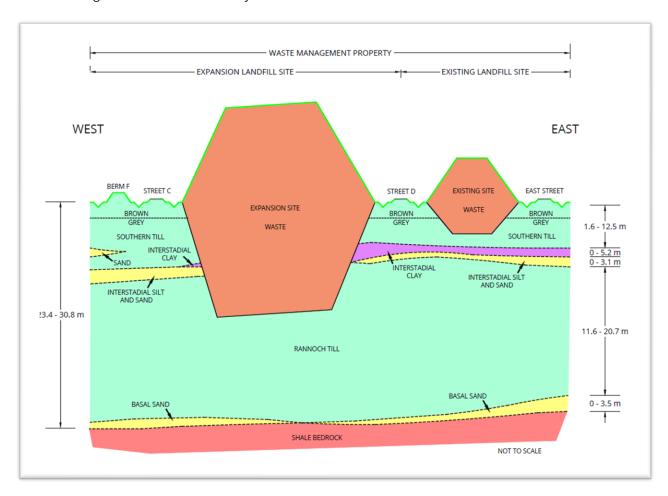
### 1.4 Site Hydrogeologic Setting

The landfill is located in southwestern Ontario within the southeastern portion of the Horseshoe Moraines physiographic region (Chapman and Putnam, 1984). The Horseshoe Moraines consist of a large horseshoe-shaped landform that includes a series of moraines aligned roughly parallel to the Lake Huron shoreline. Bear Creek to the west and Brown Creek to the east provide a slightly rolling topography around the Site. This rolling topography is further enhanced by the local tributaries and drainage swales.

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For the purpose of this report, the local stratigraphy is subdivided into the following main units: 1) the Southern Till; 2) interstadial deposits; 3) the Rannoch Till; and 4) bedrock and the overlying basal sand. Each unit is shown in the following cross-section and is briefly discussed below.



The **Southern Till** unit consists of silty clay to clayey silt with trace amounts of disseminated sand and gravel. Occasional discontinuous layers or lenses of sand were observed within this unit. At the Site, this unit ranges in thickness from approximately 1.6 m to 12.5 m. The upper 1.6 m to 5.0 m of the Southern Till is weathered and is generally identified by a brown colour. Soil fractures are numerous near ground surface and decrease in frequency with depth. The underlying unweathered Southern Till is grey in colour and contains occasional fractures.

The **interstadial deposits** generally consist of two (2) distinct deposits: 1) an upper deposit of silt and clay, often varved, that is up to 5.2 m in thickness; and 2) a lower deposit of silt to sand that is up to 3.1 m in thickness. At some borehole locations, one or both deposits were not detected which indicates that these deposits are not continuous below the Site. Below the Site, the interstadial silt and sand ranges from 4.0 m to 10.7 m below ground surface where detected.

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At about 4.0 m to 10.3 m below ground surface is the **Rannoch Till**. This till is a gritty to moderately stony clayey silt to silt till, although some textural variations occur. At some borehole locations, layers or lenses of silt to sand were detected within the till. The Rannoch Till was up to about 21.4 m thick below the Site.

Occasional discontinuous layers of sand and gravel between the Rannoch Till and underlying bedrock constitute the **basal sand**. The **bedrock** was encountered at a depth of between 23.4 m and 30.8 m below ground surface at the landfill and consists of the black bituminous shale of the Kettle Point Formation. The upper fractured portion of the bedrock and the overlying basal sand form the local bedrock aguifer.

The two (2) creeks that influence the physiography of the regional study area include Bear Creek to the west and Brown Creek to the east. The watershed for Bear Creek includes most of the Site and the area to the west. The Drainage Divide for the Bear Creek watershed cuts through the northeastern corner of the Site. Bear Creek is a tributary of the North Sydenham River and flows southwesterly to join the Sydenham River at Wallaceburg.

Surface water flow is ephemeral, with flow that typically occurs after snowmelt or prolonged periods of precipitation. Runoff from most of the east half of the Existing Site flows to Sedimentation Pond 1 (SP1), while runoff from the northern portion of Cell 11 and the west half of the Existing Site from north of approximately Cell 8/6 boundary flows to Sedimentation Pond 4 (SP4), drainage from most of Cell 6 flows into a Temporary Water Storage Area, while runoff from the remainder the west half of the Existing Site is directed to Sedimentation Pond 2 (SP2).

Sedimentation Ponds 1 through 4 are four (4) on-Site Sedimentation Ponds constructed in 2009 to manage surface water for the Site. Sedimentation Pond 2 discharges through culverts to the western Site boundary and into a tributary of Bear Creek on the east side of Lambton Road 79 (Nauvoo Road).

Surface water runoff is managed through Sedimentation Ponds 1, 2, 3, and 4. Surface water runoff originating from areas south of the Existing Site flows toward: 1) Kersey Drain (Brown Creek) to the east; and 2) to the west toward the Van Kessel Drain where it enters a municipal drainage tile at a catch basin (surface water monitoring station SS1), which is situated 60 m east of the western Site boundary. The municipal drainage tile subsequently drains into the discharge ditch of Sedimentation Pond 2, which ultimately flows westward to the Gilliand-Geerts Drain 'A', beneath Lambton Road 79.

The Brown Creek watershed drains the area east of the Site, including the northeast corner of the Existing Site. Brown Creek is a southwesterly flowing headwater of the Sydenham River, which it intersects about 1.3 kilometres (km) northeast of Alvinston.

Construction of the Sedimentation Pond network in the Expansion Site was completed by August 2009. The surface watercourse on the Expansion Site eventually discharges into a tributary of Bear Creek on the east side of Lambton Road 79 (Nauvoo Road). The surface water system around the Site is presented on **Figure 3**.

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### 1.5 Water Budget

Water budgets are provided in **Tables C-1** through **C-6**, **Appendix C**, for the 30-year normals (1961-1990, 1971-2000, 1981-2010, and 1991-2020) and the 2021 through 2022 climatic data for the area around the Site. The water budget information is based on the Thornthwaite Analytical Method (1957). A summary of precipitation data for the 30-Year Normal (1961-1990, 1971-2000, 1981-2010, and 1991-2020) and the 1995 through 2022 annual climatic data for the area around the Site is provided in **Table C-8**, **Appendix C**. The 30-year normal and data to 1996 were collected at the Strathroy Climatological Station. Data from 1997 onward were collected from the Strathroy-Mullifarry Climatological Station, which is the nearest Environment Canada climatological station to the Site. Precipitation data collected from the on-Site climatological station from January 1 to December 31, 2022, is also provided in **Table C-7**, **Appendix C**. As presented in **Table C-8**, **Appendix C**, a total of about 634.8 mm of precipitation was recorded from the on-Site climatological station during 2022, while the Strathroy-Mullifarry Climatological Station recorded about 747.1 mm of precipitation in 2022.

Relative to the 30-Year Normal (1991-2020), 2022 was drier than normal as recorded at the climatological station. The 2002 to 2022 on-Site precipitation data from January 1 to December 31 indicates that the yearly precipitation received at the Site was consistently less than the regional total. For example, the precipitation recorded from the on-Site climatological in 2022 was approximately 15.0 % less than what was measured at the Strathroy-Mullifarry Climatological Station. This pattern of annually less precipitation recorded at the Site than recorded regionally, has typically been observed since on-Site precipitation monitoring began in 2003 (2003 was first full year of monitoring).

Based on the available historical data from the Environment Canada climatological stations, there is typically a water deficit (evapotranspiration exceeds precipitation) from May through September as expected during warmer climate. Therefore, there is a low potential for infiltration (lower accumulation of groundwater) or overland flow during this period. For January through April and from September through December in 2022, a water surplus (precipitation exceeds evapotranspiration) results in a greater potential for infiltration and overland flow.

### 1.6 Monitoring System & Schedule

**Table 1** provides a schedule of the monitoring tasks completed in 2022. Also, provided in **Table 1** are rationales for monitoring requirements that were not completed in 2022. The annual monitoring program completed for the 2022 calendar year, including the relevant monitoring locations, parameters, and frequency, is outlined in **Table B-2**, **Appendix B**. **Table B-3**, **Appendix B**, provides a summary of changes to the Environmental Monitoring Plan (EMP) as approved by the Ministry of the Environment, Conservation and Parks (MECP). **Table B-4**, **Appendix B**, provides a summary of Compliance Point trigger concentration exceedances in 2022. Borehole logs and monitoring well information are provided in **Appendix D**. Monitoring well construction details are also summarized in tabular format as provided in **Table F-1**, **Appendix F**.

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### 2 MONITORING METHODS

The 2022 Compliance Monitoring Program for groundwater, surface water, landfill gas, leachate, air quality and noise were completed by RWDI between January 1 and December 31, 2022. Liquid level monitoring of the PDL and SDL of the Expansion Site was recorded automatically with pressure transducers and recorded by a SCADA system. Monitoring locations included in the annual monitoring program are listed in **Table 1** and presented on **Figure 2**.

Copies of the Chain of Custody Forms, as it relates to water, leachate, and soil, for samples submitted to the laboratory in 2022 are provided in **Table B-5**, **Appendix B**. A quality assurance and quality control (QA/QC) program was followed for each of the routine monitoring tasks completed. This program consists of procedures for the sampling of monitoring wells, the collection of samples, and ancillary tasks. A copy of the field sampling protocols is provided in **Appendix E**.

Monitoring well construction details are provided in **Table F-1**, **Appendix F**. Monitoring wells, which were included in the 2022 annual monitoring program, are designated as 'Active'. Other monitoring wells not included in the annual monitoring program are designated as 'Inactive'. Some historical monitoring wells that were recently removed and/or replaced from the monitoring program may be designated as 'Decommissioned'.

### 2.1 Liquid Level Measurements

Groundwater and leachate levels were manually measured at accessible monitoring wells, leachate wells, and maintenance holes located on the Site on May 2 and November 1, 2022.

Groundwater monitoring well nests OW82, OW83, and OW84 were installed in June 2022. Liquid levels at these well nests were measured on November 1, 2022.

Daily, during landfill operations, liquid levels for the leachate within the PDL at PS1, PS3, PS5, and PS7 (installed September 2022), as well as monthly groundwater levels for the water in the SDL at PS2, PS4, PS6, and PS8 (installed September 2022). Groundwater and leachate elevations are discussed in **Section 4**.

### 2.2 Leachate Sampling

Leachate samples were collected using dedicated disposable bailers for the relevant monitoring locations. Leachate samples were collected from select maintenance holes and pumping stations on May 6 and 10, 2022.

Leachate from the Equalization Tank was conveyed into 20 litre (L) pails, which were rinsed with leachate prior to sample collection. The sampling point for the Equalization Tank is at the truck loading bay for PS10. The Equalization Tank is sampled on a quarterly basis. Up till mid-September, the leachate within the Equalization Tank represented leachate from: PS1 (Cell 1), PS3 (Cell 2), PS5 (Cell 4), most of the Existing Site, as well as partially derived from the condensate from the landfill gas collection systems installed in the Existing and Expansion Sites. Since mid-September the leachate within the Equalization Tank represented leachate from the aforementioned sources as well as from PS7 (Cell 6A). During 2017, each cell with a leachate collection system within the Existing

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Site was automated for leachate transfer to the Equalization Tank, with the exception of the eastern portion of Cell 3S (MH3SA and MH3SB), and the northern portion of Cell 4 (MH4B). As such, leachate sampled from the Equalization Tank included leachate from the Existing Site throughout 2022.

Leachate was collected from PS1 (Cell 1), PS3 (Cell 2), and PS5 (Cell 4) via dedicated Waterra tubing equipped with a manual inertial-lift pump during the required annual sampling event in May. Leachate samples for soluble metals did not require field-filtering, but were preserved as required, while leachate samples for dissolved organic carbon (DOC) were field filtered and preserved.

Leachate chemical results for the aforementioned various sampling locations are discussed in **Section 5.1**.

### 2.3 Monitoring Well Sampling

Groundwater samples were collected from May 3 to 6, 2022 for the spring semi-annual monitoring event. During the fall semi-annual monitoring event, groundwater sampling was completed from November 1 to 2, 2022.

Groundwater samples were collected using dedicated low flow bladder pumps.

The cemetery well is sampled annually in the spring by manually purging approximately 100 L using the existing old-water supply well manual inertial-lift pump. As the water from the cemetery well could potentially be utilized as drinking water, metals sampling does not require field filtering.

As part of Site operations and to protect groundwater resources, some inactive monitoring wells were retained for potential future use as part of the monitoring program of the Expansion Site. This includes monitoring well OW59-10. Although monitoring well OW59-10 is considered an 'inactive' monitoring well, liquid levels are being assessed at OW59-10 to evaluate groundwater conditions near Cell 7 of the Existing Site. Idle monitoring wells OW39-6 and OW39-12 were decommissioned in 2017 during replacement activities related to damaged OW39-26. Monitoring locations OW61, OW62, OW75, OW76, OW77, OW78, and OW85 are currently inactive, as these monitoring wells are utilized to assess groundwater quality as it pertains to the operation of the Poplar Plantation. Since the Poplar Plantation has been inactive since its construction in 2009, groundwater quality assessment is not required to be completed in that location, which is also approved under Notice No.1 of the Sewage ECA.

Groundwater monitoring well nests OW82, OW83, and OW84 will be sampled beginning in May 2023.

Groundwater chemical results for the monitoring wells that were required to be sampled in 2022 are discussed in **Section 5.2**.

In consultation with the Landfill Engineer and Hydrogeologist Reviewers of the Technical Review Team (TRT), WM had agreed post-2016 to supplement the existing leachate level monitoring. The supplemental leachate level monitoring locations (LW1 through LW6) were installed within the Existing Site in late 2017. The locations of the select landfill gas extraction wells from within the Expansion Site were determined in 2018 [EV229 in Cell 1A (Stage 1), EV268 in Cell 1A (Stage 2), EV022 in Cell 2B, and EV226 in Cell 2D], as shown on **Figure 2**. These

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supplemental leachate level monitoring wells are proposed to be monitored semi-annually together with the already established semi-annual spring and fall monitoring events for the Site to provide further insight toward leachate patterns within select waste cells of the Existing Site, as well as the Expansion Site.

### 2.4 Surface Water Sampling

#### 2.4.1 Surface Water Flow

Surface water flow at the monitoring stations is precipitation dependent. Adequate flowing conditions to conduct surface water sampling were noted during Q1, Q2, Q3 and Q4 of 2022. After some precipitation events, some of the sampling stations did not produce the required flow for sampling. The observation of no flowing conditions after rain events  $\geq$  10mm/24hrs is expected and has been identified since precipitation monitoring began in 2003 at the Site. Verification surface water sampling events were also conducted, where required, in 2022.

As approval to discontinue monitoring the Poplar Plantation was received from the MECP on February 20, 2013, the surface water monitoring program for SS17A/B and SS18A/B remained inactive during the 2022 monitoring period.

### 2.4.2 Mechanism for Response Routine Monitoring

The surface water monitoring program adheres to the relevant Waste ECA, the Sewage ECA, as well as conditions stipulated within a MECP letter titled "RE: Request for Modification to Surface Water Monitoring/Assessment Process at Twin Creeks Landfill", dated February 27, 2014 (2014 MECP Letter). In general, the triggering mechanism for response monitoring is the receipt of  $\geq 10$  mm of precipitation in a 24-hour period from 08:00 to 08:00 hours. Flowing conditions permitting, the surface water stations are sampled within a 24-hour window following the receipt of  $\geq 10$  mm at the Site. Surface water is sampled once per quarter as part of the required routine monitoring for the Site. If a compliance monitoring station does not produce sufficient runoff to conduct sampling (or is dry), then the next precipitation event of  $\geq 10$  mm in a 24-hour period will trigger the assessment of those stations that were dry during the previous assessment. This response monitoring will continue throughout a quarterly monitoring period until the compliance stations have been sampled, or the quarter ends, in which monitoring of all the compliance stations will commence again for the next quarterly monitoring period.

### 2.4.3 Mechanism for Response Verification Monitoring

There are two (2) types of surface water monitoring stations at the Site. The first type of station consists of an open ditch where surface water flowing conditions heavily rely on precipitation. The second type of station consists of Sedimentation Ponds. When a Primary Leachate Indicator List (PLIL) parameter exceeds its trigger concentration at a surface water monitoring station, a verification monitoring response is triggered for that monitoring station.

For the 'ditch' type monitoring stations (e.g., SS1), verification monitoring can only occur after the receipt of  $\geq 10$  mm of precipitation in a 24-hour period where the precipitation was sufficient to generate flowing conditions to conduct sampling. Response monitoring continues throughout the quarter until a precipitation event is sufficient

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to conduct sampling. If insufficient flowing conditions continue at the ditch type station throughout a quarterly monitoring period, the verification event is postponed to the next quarterly monitoring period, and consequently, the postponed, monitoring event will consider both the routine quarterly monitoring event, as well as the verification monitoring event.

Similarly, for the Sedimentation Ponds, a verification monitoring event is triggered when one or more PLIL parameters exceed their relevant trigger concentrations following the routine quarterly sampling event. The verification monitoring for Sedimentation Ponds is completed within one (1) week of receipt of the routine monitoring event's chemical analytical results, independent of precipitation occurrences (Condition 5.4.) of the Sewage ECA).

Precipitation event monitoring, including biomonitoring testing is completed as outlined in the Waste and Sewage ECAs, as well as the conditions approved in the 2014 MECP Letter. An exceedance of a trigger concentration at one of the surface water monitoring compliance points would initiate verification monitoring, and if warranted, corrective action.

### 2.4.4 Precipitation Summary for Monitoring Events

#### **Monitoring Station Locations and Sampling Details:**

The surface water monitoring stations that formed part of the 2022 Compliance and Poplar System Monitoring Programs are summarized below.

Task	Monitoring Station Designations	Monitoring Station Description
	SS1	Downstream of landfill on WM property, 60 m east of Lambton Road 79 (Nauvoo Road) – Compliance Point
Surface Water	SS10	Off-Site flow into East Ditch of the Existing Site – Background Surface Water Quality
Environmental Monitoring	SS16	Flow onto expansion lands from Township land located south of the Site – Background Surface Water Quality
Program	SP1	Outlet Weir of Sedimentation Pond 1 – Internal assessment location
	SP2	Outlet Weir of Sedimentation Pond 2 – Compliance Point
	SP3	Outlet Weir of Sedimentation Pond 3 – Compliance Point
	SP4	Outlet Weir of Sedimentation Pond 4 – Compliance Point
Surface Water	SS14A (former SS14)	On-Site flow within East Ditch of the Existing Site, upstream of Poplar System
Poplar System Monitoring	SS14B (former SS15)	On-Site flow within West Ditch of the Existing Site, downstream of Poplar System as of June 2011.
Program	SS15A	South Ditch of the Existing Site and inlet point to Sedimentation Pond 1. Downstream of Poplar System.

**Note:** Former surface water monitoring stations SS14 and SS15, and revised/new surface water stations SS14A, SS14B, and SS15A as noted above, are required under the Waste and Sewage ECA's as part of the Poplar System Monitoring Program.

Details of the findings for the 2022 Poplar System surface water monitoring stations (SS14A, SS14B, and SS15A) can be found in **Volume 3** of the 2022 Annual Report.

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A summary of the precipitation that triggered the surface water monitoring events is provided below. The precipitation amounts for the five days preceding the monitoring event are also provided. The data presented in the summary reflects that collected from the on-Site climatological station in 2022.

Quarter	Previous 5 Days of Precipitation (mm)	Sampling Events	
1	0, 0, 0, 0, 20.8	February 17, 2022 – Routine monitoring event for the February 16, 2022 precipitation event.	
2	0, 5.4, 1.0, 0, 17.6	May 4, 2022 – Routine monitoring event for the May 3, 2022 precipitation event. Also, verification monitoring event for the February 17, 2022 routine precipitation monitoring event from Q1.	
2	0, 0, 0, 0, 14.4	May 16, 2022 – Verification monitoring event for the May 4, 2022 routine precipitation monitoring event from Q2.	
3	0, 0, 7.2, 0.2, 19.4	July 20, 2022 – Routine monitoring for the July 19, 2022 precipitation event.	
3	0, 0, 0, 0, 22.8	August 2, 2022 – Routine monitoring for the August 1, 2022 precipitation event. Also, verification monitoring event for the July 20, 2022 routine precipitation monitoring event from Q3.	
3	0, 0, 22.8, 0, 26.0	August 4, 2022 – Routine & Storm monitoring for the August 3, 2022 precipitation event.	
3	0, 0, 0, 2.0, 0.2*	August 30, 2022 – Routine monitoring for the August 29, 2022 precipitation event.	
3	0, 2.0, 0, 13.4, 39.0	September 28, 2022 – Routine & Storm monitoring for the September 27, 2022	
4	0.2, 0.2, 0, 0, 15.8	October 18, 2022 – Routine monitoring for the October 17, 2022 precipitation event.	
4	0.2, 0, 0, 15.8, 19.0	October 19, 2022 – Routine monitoring for the October 18, 2022 precipitation event.	
4	0, 0.6, 0, 0, 14.4	November 28, 2022 – Verification monitoring event for the October 18, 2022 routine precipitation monitoring event from Q4.	
4	N/A	January 13, 2023 – Verification monitoring event for the October 18, 2022 routine precipitation monitoring event from Q4.	

Note:

- 1) N/A denotes verification monitoring event took place at a sedimentation pond and was therefore not precipitation event dependent.
- 2) * denotes that the 24-hour precipitation total presented is not accurate as a power outage occurred for a 4-hour period resulting in the full 24-hour precipitation total not being recorded for a precipitation event that was visually observed to be greater than 10 mm/24 hours.

#### **Summary of Stations Monitored:**

A summary of the surface water stations that were sampled in 2022, including a brief explanation as to why a station was not sampled, is provided below.

Surface Water Monitoring Station	Routine Monitoring	Verification Monitoring	Rationale for No Samples Being Collected
SS1	Q1, Q2, Q3, Q4	Q1, Q2, Q4	
SS10	Q2		No flow in Q1, Q3, and Q4
SS16	Q1, Q2		No flow in Q3, and Q4
SP1	Q1, Q2, Q3, Q4	Q3	
SP2	Q1, Q2, Q3, Q4	Q2, Q3, Q4	
SP3	Q1, Q2, Q3, Q4		

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SP4	Q1, Q2, Q4	No flow in Q3
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Surface water chemical results for the aforementioned various sampling locations are discussed in Section 5.3.

#### **Biological Monitoring:**

The annual biomonitoring program was completed for the May 4, 2022, spring surface water monitoring event per the EMP. Surface water samples were collected at stations SS1, SS10, SS16, SP1, SP2, SP3, and SP4 for the annual biomonitoring program.

Verification biomonitoring was also conducted at monitoring stations SS1 on May 4 (for Q1), May 16 (for Q2), and November 28, 2022 (for Q4), at SP1 on August 2, 2022 (for Q3), and at SP2 on May 16 (for Q2), and August 2 (for Q3), 2022 as well as January 13, 2023 (for Q4) in accordance with conditions approved in 2014 MECP Letter.

Details of the biological monitoring completed during the 2022 year are discussed in Section 5.3.3 and 5.3.4.

### 2.5 Landfill Gas Monitoring

Landfill gas monitoring was completed at gas probes GP1 to GP8 in January, February, March, April, July, November, and December 2022 per the EMP. Landfill gas monitoring was completed at gas probes GP9 and GP10 in July, November, and December 2022 per the EMP, after they were installed in June 2022.

Methane gas monitoring findings are discussed in Section 5.4.

### 2.6 Automobile Shredder Residue Monitoring

Automobile Shredder Residue (ASR) may be used at the Site as daily cover in accordance with Condition 6.51 of the Waste ECA. ASR samples, when this material is utilized, are collected on a semi-annual basis during the spring and fall and submitted for analysis of Ontario Regulation 347 (O.Reg. 347) Schedule IV criteria in accordance with Condition 6.52 of the Waste ECA. Composite samples of stockpiled ASR material were collected on April 6 and October 5, 2022.

ASR sampling was not required to be completed during the 2022 Q1 monitoring period. However, a composite sample of stockpiled ASR material that was newly sourced, was collected for laboratory testing on February 2, 2022.

ASR monitoring findings are discussed in **Section 6.3**.

### 2.7 Contaminated Soil Monitoring

Contaminated soil may be managed on-Site in accordance with Conditions 6.53 to 6.61 of the Waste ECA. Quarterly sampling is completed for contaminated soil that is utilized as daily and/or intermediate cover for the Expansion Site per Conditions 6.56 and 6.57 of the Waste ECA, and is submitted for analysis of O.Reg. 347 Schedule IV. Composite samples of contaminated soil material were collected on January 11 (for Q1), April 6 (for Q2), July 6 (for Q3), and October 5 (for Q4) in 2022.

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Contaminated soil monitoring findings are discussed in Section 6.4.

### 2.8 Field Sampling Parameters

Groundwater, leachate, and surface water field testing included the measurement of pH, electrical conductivity (EC), temperature, and turbidity (groundwater, surface water, and leachate), as well as dissolved oxygen (DO) (for surface water only). Surface water field parameters, including flow rates, were measured at each monitoring station, when water was present, during each monitoring event.

### 2.9 Laboratory Analytical Parameters

Analytical parameters are listed in **Table B-2**, **Appendix B**. In general, analyses were completed by Bureau Veritas Laboratories, a Canadian Association of Laboratory Accreditation (CALA) accredited laboratory. Analyses for Contaminated Soil monitoring in Q1 were completed by Eurofins Scientific, which is also a CALA accredited laboratory.

### 2.10 Field QA/QC Sampling

The field sampling QA/QC program is outlined in the following table. No equipment rinse blanks were collected due to the use of dedicated sampling equipment.

Media	Monitoring Event	Field-prepared Duplicate (Original Sample)
Leachate	May 6, 2022	LDUP (MH18)
		GWDUP1 (OW67-11)
	May 2 to 6, 2022	GWDUP2 (OW68-5)
	May 3 to 6, 2022	GWDUP3 (OW80-3)
Groundwater		GWDUP4 (OW40A-7)
		GWDUP1 (OW70B-5)
	November 1 to 2, 2022	GWDUP2 (OW16-6)
		GWDUP3 (OW16-7)
	Fabruary 17, 2022	SSDUP1 (SS1)
	February 17, 2022	SPDUP (SP1)
	May 4 2022	SSDUP1 (SS1)
Surface Water	May 4, 2022	SPDUP (SP4)
Surrace water	luly 20, 2022	SSDUP1 (SS1)
	July 20, 2022	SPDUP (SP2)
	October 19, 2022	SSDUP1 (SS1)
	October 18, 2022	SPDUP (SP1)

**Notes:** 1) Field and trip blanks were analyzed as part of the groundwater monitoring events during May and November 2022. Parameters are outlined in **Table B-2**, **Appendix B**.

²⁾ Field prepared duplicate samples are not required for verification monitoring events.

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### 3 QUALITY ASSURANCE & QUALITY CONTROL EVALUATION

The QA/QC program included field-prepared duplicate samples, field-prepared blanks, trip blanks, comparisons with field-determined analytical results, laboratory-prepared blanks, matrix spikes, duplicates, percent recoveries of analyses, and data review. Additionally, the QA/QC program consists of verifying that the correct parameters were assessed by the laboratory for each sampling event, and that they were assessed within their respective hold times (not beyond the sample expiration date). During 2022 the correct parameters were assessed within the allotted hold times.

The laboratory analyzed several control samples to verify that the analytical equipment was functioning properly and reporting results accurately at the time of analysis for the samples collected at the Site. The control samples had an expected target value, which was compared against pre-determined data quality objectives. For the laboratory control samples, the results were within acceptable laboratory data quality criteria.

For the field-prepared duplicate samples, the analytical results for the required parameters of analysis, completed as detailed in **Section 2.9**, were evaluated for the relative percent difference (RPD) of parameter concentrations using the United States Environmental Protection Agency (USEPA) National Functional Guidelines (US EPA 542-R-20-006 and US EPA 540-R-20-005) as a general QA/QC RPD screening mechanism. The RPD screening mechanism is such that for concentrations greater than five (5) times the laboratory reportable detection limit (RDL), a concentration difference of less than or equal to 20% is deemed acceptable. For concentrations less than or equal to five (5) times the RDL, a concentration difference of equal to or less than the RDL is deemed acceptable. Where an exceedance of the general QA/QC RPD screening mechanism is identified, the results for the required parameters of analysis are evaluated against the applicable performance standards for sample duplicates noted in Tables 5.1 to 5.15 of the *Protocol for Analytical Methods Used in the Assessment of Properties under Part XV.1 of the Environmental Protection Act*, as prepared by the Ministry of the Environment, Conservation, and Parks (MECP), dated March 9, 2004, and amended on February 19, 2021. For the results found to exceed the criteria of each QA/QC evaluation, a laboratory data quality review (DQR) of the results is completed by the laboratory to verify that the concentrations are accurate as presented and are within acceptable laboratory data quality criteria.

Laboratory chemical results for the leachate, groundwater, and surface water are provided in **Appendices G**, **H** and **I**, respectively.

### 3.1 Leachate

For leachate samples collected for the 2022 Compliance Monitoring Program, the RPD were acceptable between original and duplicate samples. Therefore, the results for the leachate samples collected during the 2022 monitoring event were considered representative of actual leachate quality at the time of sample collection and were acceptable for inclusion into the database for interpretation.

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The leachate field analytical results for temperature, pH, EC, turbidity, and DO, are provided for the Existing Site (CFA-Comp, Sump, and MH18), as well as the Expansion Site (PS1, PS3, PS5, and Equalization Tank) in **Table G-1**, **Appendix G**. Field leachate temperatures showed some variability reflective of the location the leachate is being stored/generated and the time of year the leachate was assessed. For example, the quarterly field temperature results for the Equalization Tank showed expected variability based on the time of year sampled with relatively lower temperatures for January and November, and higher temperatures for May and July.

The field analytical values recorded for pH across the Site varied between 7.1 and 8.4 pH units. EC values also varied, with a range of 1,650 to 3,730 micro-Siemens per centimetre ( $\mu$ S/cm) at the Existing Site, and 10,580 to >20,000  $\mu$ S/cm for the Expansion Site. Turbidity values also expectedly varied with values between 17.9 and 213 nephelometric turbidity units (NTU) for the Existing Site, and values between 105 and >1,000 NTU for the Expansion Site.

Ion balances were also calculated as a QA/QC procedure for the leachate. Considering major anions and cations, an ion balance difference of greater than 10% would initiate a more thorough review of the chemical results and laboratory procedures. The leachate chemical results obtained during the 2022 annual monitoring program satisfied the 10% ion balance target, except as noted in the summary below. The cause of the ion balance exceedances summarized below are reasonable for the noted parameters and the respective concentrations detected in the relevant samples.

Media	Station/Monitoring Well ID	Ion Balance Percentage and Date	Comments
	CFA-COMP	15.6 – May 2022	High anion concentrations
	Equalization Tank	26.1 – May 2022	High anion concentrations
Leachate	Equalization rank	14.1 – November 2022	High anion concentrations
Leachate	PS1	39.1 – May 2022	High anion concentrations
	PS3	25.0 – May 2022	High anion concentrations
	PS5	11.5 – May 2022	High anion concentrations

### 3.2 Groundwater

For the groundwater samples collected for the 2022 Compliance Monitoring Program, the RPD were acceptable between original and duplicate samples, with the exception of select parameters as summarized below.

OW67-11 (GWDUP1)  Total Ammonia Dissolved Iron  Conductivity Total Dissolved Solids Dissolved Sulphate Alkalinity	OW67-11 (GWDUP1)  Dissolved Iron  Conductivity  Total Dissolved Solids  Dissolved Sulphate	OW67-11 (GWDUP1)  Dissolved Iron  Conductivity  Total Dissolved Solids  Dissolved Sulphate  Alkalinity	Sampling Date	Original Sample (Duplicate)	Parameter with QA/QC Exception
May 3 to 6, 2022  Total Dissolved Solids  Dissolved Sulphate  Alkalinity	May 3 to 6, 2022  OW68-5 (GWDLIP2)  Total Dissolved Solids Dissolved Sulphate Alkalinity	May 3 to 6, 2022  OW68-5 (GWDUP2)  Total Dissolved Solids Dissolved Sulphate Alkalinity Dissolved Chloride		OW67-11 (GWDUP1)	
May 3 to 6, 2022  Dissolved Sulphate Alkalinity	May 3 to 6, 2022  OW68-5 (GWDLIP2)  Dissolved Sulphate Alkalinity	May 3 to 6, 2022  OW68-5 (GWDUP2)  Dissolved Sulphate Alkalinity Dissolved Chloride			,
Alkalinity	OW68-5 (GWDUP2)  Alkalinity	OW68-5 (GWDUP2)  Alkalinity  Dissolved Chloride	May 3 to 6, 2022		
	Dissolved Chloride	Dissolved Chloride		OW68-5 (GWDLIP2)	Alkalinity
Dissolved Barium Dissolved Boron	Dissolved Boron				Dissolved Calcium

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		Dissolved Magnesium
		Dissolved Potassium
		Dissolved Sodium
	OW80-3 (GWDUP3)	Dissolved Potassium
		Dissolved Sodium
	OW40A-7 (GWDUP4)	Dissolved Organic Carbon
		Dissolved Iron
November 1 to 2, 2022	OW70B-5 (GWDUP1)	Total Dissolved Solids

A laboratory DQR for each of the aforementioned results indicated that the concentrations for the above noted parameters were accurate as presented and within acceptable laboratory data quality criteria. Therefore, the results for the leachate samples collected during the 2022 monitoring event were considered representative of actual leachate quality at the time of sample collection and were acceptable for inclusion into the database for interpretation.

The groundwater field analytical results for temperature, pH, EC, and turbidity are provided for the active aquitard, the interstadial silt and sand, as well as the interface aquifer in **Tables H-1** to **H-3**, **Appendix H**, respectively. Shallow groundwater generally showed greater temperatures than the deeper groundwater during May and cooler temperatures compared to the deeper groundwater during November, indicative of seasonal influences on the shallower groundwater. The field analytical values recorded for pH varied between 6.5 and 8.2 pH units. As expected in clayey soil, both conductivity and turbidity values varied, with a range of 860 to 5,220  $\mu$ S/cm for conductivity and 1.05 to >1,000 NTU for turbidity.

Ion balances were also calculated as a QA/QC procedure for the groundwater. Considering major anions and cations, an ion balance difference of greater than 10% would initiate a more thorough review of the chemical results and laboratory procedures. The groundwater chemical results obtained during the 2022 annual monitoring program satisfied the 10% ion balance target, except as noted in the summary below. The cause of the ion balance exceedances summarized below are reasonable for the noted parameters and the respective concentrations detected in the relevant samples.

Media	Station/Monitoring Well ID	Ion Balance Percentage and Date	Comments
	OW17-4	30.6 – May 2022	High cation concentrations
		31.5 – November 2022	High cation concentrations
Groundwater	OW40D-4	15.0 – May 2022	High cation concentrations
Groundwater	OW70B-5	11.1 – November 2022	High cation concentrations
	OW80-3	17.3 – May 2022	High cation concentrations
	Cemetery Well	10.2 – May 2022	High cation concentrations

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### 3.3 Surface Water

For the surface water samples collected for the 2022 Compliance Monitoring Program, the RPD were acceptable between original and duplicate samples, with the exception of select parameters as summarized below.

Sampling Date	Original Sample (Duplicate)	Parameter with QA/QC Exception
February 17, 2022	SS1 (SPDUP1)	Total Arsenic Total Iron
	SS1 (SSDUP1)	Total Phosphorus
May 4, 2022	SP4 (SPDUP)	Total Chemical Oxygen Demand Dissolved Sulphate
August 8, 2022	SS1 (SSDUP1)	Total Suspended Solids
	SP2 (SPDUP)	Total Suspended Solids
October 18, 2022	SS1 (SSDUP1)	Nitrite Total Cadmium

A laboratory DQR for each of the aforementioned results indicated that the concentrations for the above noted parameters with RPDs were accurate as presented and within acceptable laboratory data quality criteria. Therefore, results for the surface water samples collected during the 2022 monitoring events were considered representative of actual groundwater quality at the time of sample collection and were acceptable for inclusion into the database for interpretation.

Surface water field analytical results are provided in **Table I-1**, **Appendix I**. During the 2022 monitoring events, the surface water temperatures reflected the ambient air temperature during sampling. Field pH, EC, turbidity, and DO values fluctuated with no notable anomalies.

Ion balances were also calculated as a QA/QC procedure for surface water. Considering major anions and cations, an ion balance difference of greater than 10% would initiate a thorough review of the chemical results and laboratory procedures. The surface water chemical results obtained during the 2022 annual monitoring program satisfied the 10% ion balance target, except as noted in the summary below. The cause of the ion balance exceedances summarized below are reasonable for the noted parameters and the respective concentrations detected in the relevant samples.

Media	Surface Water Monitoring Station	Ion Balance Percentage and Date	Comments
		15.1 – February 2022	Low anion concentrations
Surface Water		21.7 – October 2022	Low anion concentrations
	SP1	23.3 – February 2022	Low anion concentrations

In summary, the 2022 field and laboratory QA/QC results indicated that the chemical results were representative of actual conditions at the time of sample collection.

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## 4 GROUNDWATER & LEACHATE ELEVATION RESULTS

To define the local groundwater setting below the landfill, the stratigraphic sequence described in **Section 1.4** can be associated with hydrostratigraphic units, as summarized below.

Stratigraphic Unit	Hydrostratigraphic Unit	Approximate Depth to top of Unit (m)
Brown Zone in the Southern Till	Active Aquitard	0.0
Grey Zone in the Southern Till	Upper Aquitard	164070
Interstadial Clay and Silt		1.6 to 7.9
Interstadial Silt and Sand	Interstadial Sand	4.0 to 10.7
Rannoch Till	Lower Aquitard	4.5 to 12.5
Fractured Bedrock and Basal Sand	Interface Aquifer	22.8 to 29.3

Although each stratigraphic unit is identified as part of a hydrostratigraphic unit, each unit has a hydraulic influence on the others. Groundwater monitoring wells developed within each hydrostratigraphic unit are summarized below and monitoring well construction details provided in **Table F-1**, **Appendix F**.

Hydrostratigraphic Unit	Monitoring Wells
Active Aquitard	OW16-6, OW17-4, OW40D-4, OW54A-4, OW56-4, OW57-4, OW58-6, OW59-6, OW60-4, <i>OW61-4, OW62-5</i> , OW67-4, OW68-5, OW69-5, OW70B-5, OW71A-5, OW72-6, OW73-6, <i>OW75-3, OW76-5, OW77-4, OW78-4,</i> OW79-5, OW80-3, OW81-5, OW82-5, OW83-5, OW84-6, <i>OW85-5</i>
Interstadial Silt and Sand	OW16-7, OW40A-7, OW46-7, OW47-6, OW54-10, OW57-15, OW58-17, OW60-8, <i>OW61-6, OW62-7</i> , OW67-11, OW72-10, OW73-9, <i>OW75-7, OW78-6,</i> OW79-7, OW80-6, OW81-7, OW82-14, OW83-9, OW84-11, <i>OW85-8</i>
Interface Aquifer	OW17-30, OW19-29, OW39A-26, OW40A-28, OW49-29, OW60-25, <i>OW61-26, OW62-30</i> , OW79-26, OW80-27, OW81-27, OW82-28, OW83-29, OW84-31, Cemetery Well

**Note:** *Italicized* denotes monitoring well is inactive as the monitoring well is to be used to evaluate the Poplar Plantation two (2) months prior to activation of the system.

Monitoring well nests OW82, OW83, and OW84 installed in June 2022.

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### 4.1 Leachate Elevations

Leachate elevations measured at the Existing Site during the 2022 annual monitoring program are presented with historical data in **Tables F-2** and **F-3**, **Appendix F**. Leachate elevation data from the pumping stations of the Expansion Site are presented with historical data in **Table F-6**, **Appendix F**.

### 4.1.1 Leachate Elevation Assessment - Existing Site

#### **Monitoring Details:**

Leachate elevations are plotted on **Figure 4**. Flow valves between maintenance holes for each cell of the Existing Site are left in a normally open position to facilitate the automated pumping of leachate to the Equalization Tank, as necessary to keep leachate levels at an acceptable elevation.

#### **Collection System(s) Background:**

To provide background for the understanding of the design of the waste cells within the Existing Site, Cells 3S, 4, 5, 6, 7, 8, 9, 10, and 11, as well as Cell 12 (only the southern third is constructed) contain waste underdrains that direct leachate to a perimeter collector system. Cells 10 and 12 are hydraulically connected, with leachate extraction typically occurring from MH12. The South Cell is completed with finger drains that direct leachate to a perimeter collection system. The West Cell is completed with a collection "Sump" to collect leachate for extraction purposes. Cell 3 does not have a leachate collection system, however, based on historical liquid levels the leachate in this cell is understood to be managed by the adjacent cells. For cells with waste underdrains, when the valve is open and leachate flows freely between a cell's maintenance holes, the leachate elevation in the maintenance holes of that cell is expected to be similar, typically within 0.15 m.

Two (2) exceptions to this pattern can occur for cells with waste underdrains, which are for Cells 3S and 4. For Cell 3S, the leachate levels can be notably different in each maintenance hole. By the original design of the collection system MH3SA and MH3SB were not hydraulically connected with a pipe and are the upper end of the system, which drain to the low end at MH3SC. As of October 2022, for operational efficiencies, MH3SB was converted into a pumping station and MH3SA and MH3SB were hydraulically connected with a pipe. As such the leachate elevations in MH3SA and MH3SB will be expected to be similar, typically within 0.15 m. MH3SC is connected with a pipe to MH3SD but is constructed with a sump ~2m deeper than the base of MH3SD. For Cell 4, there are two distinct (north third and south two-thirds) waste underdrain systems, which independently gravitationally drain to their respective low ends, MH4A (south system) and MH4B (north system). MH4B gravity drains through a toe drain to MH4A.

By the fall of 2017 each of the downstream maintenance holes for the Existing Site were updated to be operated as pumping stations. This pumping station conversion enables the leachate level in the maintenance hole to be automatically managed, year-round, via pumping leachate to either the equalization tank or to the Poplar System irrigation tanks.

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### **Leachate Management:**

As a general leachate best management practice, the extraction of leachate within the Existing Site is prioritized based on the leachate elevations for each individual waste cell to facilitate inward hydraulic gradients (as calculated considering the previous year's seasonally relevant groundwater elevations). Exceptions occur since 2019 for leachate management practices where leachate is stored in the waste during late winter to spring, for volume availability for irrigation application to the Poplar System in the subsequent growing season. From an environmental stewardship perspective, the leachate generated from within the waste mound of the Existing Site was determined to generally be of more suitable use as irrigation liquid during the growing season as opposed to dedicating efforts for off-Site disposal and treatment between late winter 2022 and spring 2022. It is noted that storing leachate in this manner is suitable as environmental compliance for the Existing Site relies on groundwater and surface water quality monitoring, which were acceptable.

#### **Leachate Level Assessment Details:**

The hydraulic assessment of leachate compared to nearby shallow groundwater is utilized as an additional guidance tool toward managing leachate volumes in the waste cells and preventing the potential for releases to local water resources, such as surface water and groundwater. The leachate target elevations updated in 2021 were used to assist in lowering the liquid elevation within the maintenance holes and Sump during 2022. Most of the waste cells of the Existing Site are completed with waste underdrains such that leachate will percolate vertically downward through the waste and into the collection lines and be directed gravitationally toward the perimeter maintenance holes. As the waste mound is not likely uniform in nature, some leachate could be perched within the waste mound. As such, the potential for leachate seeps exists. In 2022, RWDI did not observe any seeps on the landfill cap during their respective Site inspections.

#### 4.1.1.1 Leachate Elevation Patterns – Existing Site

It is apparent that leachate elevations varied across the Site. A comparison of the May and November 2022 leachate elevations with ground surface and inferred groundwater elevations outside the waste footprint is provided in **Table 2**. Overall, leachate elevations generally increased from November 2021 to May 2022 and/or then decreased from May 2022 to November 2022. This generalized pattern was expected for 2022 as discussed in more detail in the following sections. Exceptions to the generalized pattern generally were noted in the western portion of Cell 3S (MH3SC and MH3SD), Cell 10, Cell 11, and Cell 12, where leachate elevations slightly increased from May 2022 to November 2022. This was a result of a reduction in the application of leachate in 2022 compared to 2019 and 2020.

Based on 2022 leachate elevations, the hydraulically connected maintenance holes in the Existing Site generally showed an elevation difference that was equal within the same cell or that differed by less than 0.15 m. Exceptions to this occurred during one or both monitoring events for Cell 3S (at MH3SC and MH3SD), Cell 8, Cell 9, and Cell 11 in 2022. The below summary outlines the noted differences, which are not a concern as the differences are either based on design or generally consistent with historical observations.

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- Within Cell 3S, the leachate elevation at MH3SC for the May and November monitoring events were greater than the elevations at MH3SD by 0.20 m and 0.31 m, respectively, which is generally consistent with historical observations.
- Within Cell 8, the leachate elevations at MH8A for the May and November monitoring events were less than the elevations at MH8B by 0.19 m and 0.28, respectively, which is generally consistent with historical observations.
- Within Cell 9, the leachate elevation at MH9A for the May monitoring event was greater than the
  elevation at MH9B by 0.18 m, which is generally consistent with historical elevations.
   Within Cell 11, the leachate elevation at MH11A for the November monitoring event was less than the
  elevation at MH11B by 0.31 m, which is generally consistent with historical elevations.

#### 4.1.1.2 Leachate Levels Assessed Against Ground Surface

Leachate elevations/pressures that would have the potential to affect the groundwater systems are reflected by the leachate elevations within the associated maintenance holes of a given waste cell. As presented in **Table 2**, leachate elevations within the eastern and western waste cells were generally lower than the surrounding ground surface in May and November 2022. One (1) exception occurred in the West Cell (Sump) both during the May and November 2022 monitoring events. Since the November 2015 monitoring event, the leachate elevation within the Sump has generally exhibited an increasing trend until November 2019, which showed a significant decrease of 2.6 m as a result of leachate extraction. The leachate elevation within the Sump showed a further decrease from November 2019 to November 2020 of 1.51 m, also as a result of leachate extraction.

Overall, however, acceptable groundwater and generally acceptable surface water quality was noted around the Existing Site in 2022. Additionally, leachate seeps were not observed along the waste side slopes of the Existing Site in 2022. Therefore, the noted leachate elevation within the West Cell does not represent an immediate concern. Continued leachate extraction from the West Cell as well as the South Cell via automated pumping is expected to further reduce the leachate mound in these cells. Ongoing monitoring will evaluate these locations over time.

4.1.1.3 Leachate Levels Assessed Against the Shallow Groundwater Table

#### **Spring 2022:**

In May 2022, leachate elevations were lower than the local and inferred shallow groundwater table for Cells 3, 4, 5, 7, 8, 9, 10, 11, 12, Cell 3S (western portion), as well as the South Cell (eastern portion). This comparison shows that the groundwater was being induced toward the waste and that leachate was hydraulically contained within the waste at these locations.

• Within Cell 3S (at MH3SA and MH3SB), the leachate elevations were higher than the local shallow groundwater elevations to the east by 0.52 m and 0.42 m, respectively. It is noted that the May 2022 leachate elevations within MH3SA and MH3SB are within the historical ranges for these locations.

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- Within Cell 6 (at MH6A), the leachate elevation was higher than the historical local shallow groundwater
  elevation by 2.82 m. It is noted that the May 2022 leachate elevation within MH6A is within the historical
  range for this location.
- Within the West Cell (Sump), the leachate elevation was higher than the historical local shallow groundwater elevations by 5.29 m. However, the May 2022 leachate elevation within the SUMP is 2.71 m lower than the peak elevation observed in May 2019. An evaluation of the ability to further lower the leachate elevation in the Sump will be completed during 2023.
- Within the South Cell (at OW22A-10 and OW53-10) the leachate elevations were higher than the historical local shallow groundwater elevation to the west by 0.32 m, and 1.76 m, respectively. It is noted that the May 2022 leachate elevations within OW22A-10 and OW53-10 are within the historical ranges for these locations.

#### Fall 2022:

In November 2022, the leachate elevations throughout select cells of the Existing Site expectedly showed a decrease since May 2022 as a result of leachate extraction for irrigation onto the Poplar System. For a few waste cells, the leachate elevations remained elevated with respect to the historical shallow groundwater elevations to the west and south of the Existing Site. However, leachate elevations were lower than the local and inferred shallow groundwater table for Cells 3, 4 (southern portion) 5, 7, 8, 9, 10, 11, 12, Cell 3S (eastern portion), as well as the south cell (eastern portion).

- Within Cell 3S (as MH3SC), the leachate elevation was higher than the local shallow groundwater elevation to the west by 0.30 m. It is noted that the November 2022 leachate elevation within MH3SC is within the historical ranges for these locations.
- The leachate elevation within the northern portion of Cell 4 (at MH4B) was higher than the shallow groundwater elevation by 0.57 m. It is noted that the November 2022 leachate elevation within MH6A is within the historical range for this location.
- The leachate elevation within Cell 6 (at MH6A) was higher than the historical shallow groundwater elevation by 2.89 m. It is noted that the November 2022 leachate elevation within MH6A is within the historical range for this location.
- Within the West Cell (Sump), the leachate elevation was higher than the historical local shallow groundwater elevations by 5.42 m. However, the November 2022 leachate elevation within the SUMP is 2.58 m lower than the peak elevation observed in May 2019. As discussed above, an evaluation of the ability to further lower the leachate elevation in the Sump will be completed during 2023.
- Within the South Cell (at OW22A-10 and OW53-10) the leachate elevations were higher than the historical local shallow groundwater elevation by 0.23 m, and 3.06 m, respectfully. It is noted that the November 2022 leachate elevations with OW22A-10 and OW53-10 are within the historical ranges for these locations.

Overall, as discussed herein, acceptable groundwater and surface water quality was demonstrated at the compliance points during 2022. Also, leachate seeps were not observed along the waste side slopes of the Existing Site in 2022. Therefore, the noted leachate elevations that were higher in 2022 than the local

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groundwater elevations within select leachate monitoring wells and maintenance holes of the Existing Site did not represent a concern.

#### 4.1.1.4 Leachate Elevation Trends – Existing Site

Hydrographs of leachate elevations are presented on **Figures F-1** through **F-12**, **Appendix F**. Long-term trends are summarized in **Table 3**. For the purpose of assessing elevation trends, a constant elevation trend is defined as having 0.1 m or less of seasonal variation.

Over past years, leachate elevations have mostly fluctuated with some exceptions, as noted in **Table 3**. The historically increasing leachate elevation trends that were noted in 2019 for Cell 3S (at MH3SC and MH3SD), Cell 6 (at MH6A and MH6B), and the West Cell (Sump) are noted to have decreased since then or have generally been fluctuating below their historical highs. Notwithstanding this trend, an evaluation of the ability to further lower the leachate elevation in the Sump will be completed during 2023.

Over the long-term, leachate elevations are expected to continue to decrease and eventually stabilize since the Existing Site is capped and leachate is extracted from cells within the Existing Site for the operation of the Poplar System and for off-Site disposal. It should be noted that over time, the amount of leachate removed from storage will likely increase with the operation of the expanded Poplar System, as the system matures. The volume of leachate managed from the Site in 2022 is summarized in **Section 4.1.3.1**.

#### 4.1.1.5 Supplemental Leachate Level Assessment

As discussed in **Section 2.3**, supplemental leachate level monitoring wells were installed in 2017 and were monitored semi-annually in 2022 together with the already established semi-annual spring and fall monitoring events at the Site. The leachate elevations for the leachate level monitoring wells (LW1 to LW6) are presented in **Table F-3**, **Appendix F.** 

#### **Existing Site:**

The May and November 2022 data indicates that select waste cells have leachate stored within the waste that may not be directed to, or captured by, the leachate underdrain collection systems. As such, this leachate does not represent a hydraulic loading to the respective cell's floor/liner and therefore, does not have a direct influence on the groundwater system (i.e. potentiometric influences).

During 2022, the leachate elevations within select maintenance holes were generally drawn down throughout the majority of the application period and fluctuated significantly as a result of leachate extraction for irrigation to the Poplar System. With respect to draw down as a result of leachate extraction, the liquid elevations in leachate monitoring wells LW1 to LW6 did not definitively correlate to the elevations in their counterpart maintenance hole (within the same waste cell). This observation indicates that leachate within the waste mound that may not be directed to, or captured by the leachate underdrain collection system (i.e. perched, low hydraulic conductivity, or some other cause) is therefore, not having an effect to the groundwater system (i.e. potentiometric pressures). This effect is observed in waste cells that are both positioned under and not under the Poplar System, indicating that the operation of the Poplar System is not the cause of this leachate elevation differential effect.

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There was no observable negative affect observed to the sideslope of the Existing Site, such as leachate seeps, soil staining, stress vegetation, soil slumping or erosion, as a result of this leachate elevation differential. At this time the exact cause of the leachate differential is unknown, but it is not causing a detrimental effect to the operation of the Poplar System, excluding limiting leachate volume availability from select waste cells, and is not causing visible stress to the landfill cap and as such does not represent an immediate concern. Ongoing leachate level monitoring will enable an evaluation if this leachate differential is a long-term (i.e. over 5 years) situation, or has the potential to represent a concern (i.e. increasing trends with time, or stress to the landfill cap, etc.).

In summary, the Existing Site requires the ongoing leachate level monitoring program to continue to enable input such that the leachate is managed environmentally effective in consideration of the destination target (e.g. off-Site vs. on-Site treatment).

#### 4.1.1.6 Leachate Storage Volume – Existing Site

Overall, between May 2021 and May 2022 there was a slight increase in the calculated theoretical total leachate volume stored above the local groundwater table (31,904 m³) and within the waste (13,284 m³) of the Existing Site, as summarized in Table 5. Between November 2021 and November 2022 there was a decrease in the calculated theoretical total leachate volume stored above the local groundwater table (34,275 m³) and within the waste (36,911 m³) of the Existing Site, as summarized in **Table 5**. It is noted that these volumes are only theoretical in nature and are used as guidance information for year over year overall performance evaluation. For example, the total leachate storage in the Existing Site decreased from May to November 2022 (a period of time approximately 2 weeks longer than the irrigation season) by 28,572 m³, however, during that time approximately 7,173 m³ was measured to have been extracted from the Existing Site. During this timeframe, the pumping stations in many of the weaker strength leachate waste cells were routinely drawn down to elevations where the leachate elevation was too low to safely operate the pumps.

As discussed, and as presented on Figures F-1 to F-12, after the 2022 irrigation season in November 2022, the leachate elevations in a few of the cells of the Existing Site were determined to be slightly higher than the May 2022 elevations whereas, the leachate elevations in the majority of the cells were determined to be lower.

Over the long-term, leachate volumes within cells of the Existing Site are expected to continue to decrease as more leachate is extracted for the operation of the expanded Poplar System and off-Site disposal.

The calculation method or input variables for the theoretical leachate storage volume will be revisited in an effort to have more accurate theoretical values compared to actual values based on field data interpretations/expectations.

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### 4.1.2 Leachate Elevation Assessment – Expansion Site

Leachate level monitoring within the Expansion Site is completed to assess leachate pressures on the liner and the effectiveness of the leachate management system. Leachate levels for PS1, PS3, PS5, and PS7, are tabulated in **Table F-6** and graphically represented on **Figures F-26 to F-29**.

#### **Collection System Background:**

Leachate within each cell is directed to a sump where it is managed by pumping stations PS1, PS3, PS5, and PS7, (Cell 1, Cell 2, Cell 4, and Cell 6A, respectfully) for conveyance to the Equalization Tank. The operation of the pumps in the aforementioned pumping stations is SCADA-controlled (automated) with liquid level sensors that control pump 'on' and 'off' depending on the leachate level within the relevant sump. Leachate retrieval from the Equalization Tank is controlled by pump station PS10. Pumping station PS1 began operation on November 16, 2009. PS3 began operation on November 21, 2013. PS5 began operation on October 1, 2019. PS7 operation began on September 14, 2022.

#### **Trigger Mechanism Assessment Process:**

The trigger mechanism for implementation of groundwater contingency measures for the Expansion Site is the loss of hydraulic containment of the landfill waste footprint. The loss of hydraulic containment occurs when leachate levels within the PDL are higher than the surrounding groundwater elevation for the active aquitard and the groundwater pressures for the interface aquifer.

Leachate liquid levels for the pumping stations for the Expansion Site are evaluated against the Waste ECA Conditions 7.18 and 14.1.

- Condition 7.18 of the Waste ECA requires that a maximum leachate head of 300 mm (or 0.30 metres (m)) on the landfill liner is not exceeded.
- Condition 14.1 of the Waste ECA requires that, for PS1, PS3, PS5, and PS7, the trigger leachate elevations
  for hydraulic containment for the protection of groundwater are 232.7 metres above sea level (m ASL),
  232.6 m ASL, 232.8 m ASL, and 233.4 m ASL, respectively. These elevations represent a maximum
  leachate head of 6 m above the landfill liner of each respective cell.

Occasionally, after major storm events when a part of the active waste disposal area within the Expansion Site is not final capped, a large percentage of precipitation will move to the cell floor and the leachate pumping rates for PS1, PS3, PS5, and PS7 would be less than the incoming volume of liquid.

It is noted that periodic elevated leachate occurrences are described in the MECP-approved Development and Operations Report (D&O Report) (Henderson Paddon, 2008). A temporary increase of the leachate head under such circumstances is not considered a non-compliance issue with the Waste ECA. Notwithstanding this consideration, WM has implemented an Expansion Landfill Leachate Level Contingency Plan (RWDI, April 30, 2021). The aforementioned plan was implemented so that WM is aware of a leachate level elevation that could be indicative of a pending operating scenario that could cause the leachate head to exceed 300 mm. Notification process will enable WM to be aware of a level of possible concern and enable WM to evaluate if the leachate level

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is a concern or not. If the level is of concern, then WM will be able to evaluate the leachate level operational system to identify the cause of the leachate level of concern and implement the relevant remedial measures detailed in the D&O to address operational issues for leachate level management. The notifications are denoted below.

- Electronic issuance (email and/or text) to relevant WM operations personnel that a leachate level for a given waste cell of the Expansion Landfill is at 80% (240 mm) of the 300 mm leachate head limit.
- A visual notification (i.e., light) on the leachate pumping stations when the leachate level for the respective waste cell of the Expansion Landfill is at 80% (240 mm) of the 300 mm leachate head limit.

#### Condition 7.18

During 2022, the leachate elevations at PS1, PS3, PS5, and PS7 generally satisfied condition 7.18 of the Waste ECA, with the exception of one (1) occurrence at PS7 on September 29th, where the leachate elevation exceeded the 0.3 m of head by 0.01 m.

#### Condition 14.1

During 2022, the leachate elevations at PS1, PS3, PS5, and PS7 remained well below their respective trigger leachate elevations, as outlined in Condition 14.1 in the Waste ECA.

In addition, the leachate elevations at PS1, PS3, PS5, and PS7 were below the historical groundwater elevation in the active aquitard and the interstadial silt and sand flow systems. Therefore, the leachate in Cells 1, 2, 4, and 6A was hydraulically contained from the afore-mentioned groundwater flow systems.

Further details are included in **Appendix Q.** 

#### 4.1.2.1 Leachate Elevation trends – Expansion Site

Leachate levels for PS1, PS3, PS5, and PS7, are presented in **Table F-6** and on **Figure F-26 to F-29**, **Appendix F**. As expected, leachate levels vary with time as a result of the gradual accumulation of leachate and pumping activities to maintain acceptable leachate levels at each pumping station.

#### 4.1.2.2 Supplemental Leachate Level Assessment

#### **Expansion Site:**

The leachate elevations for the Expansion Site were collected from landfill gas extraction wells, otherwise known as early vertical gas wells (EVGW) [EV229 in Cell 1A (Stage 1), EV268 in Cell 1A (Stage 2), EV022 in Cell 2B, and EV226 in Cell 2D] are presented in **Table F-8, Appendix F**. The locations of these EVGWs are shown on the attached **Figure 2**.

Leachate elevations for the EV229, EV268, EV022, and EV226 were unable to be measured during 2022 semiannual monitoring events. For the EVGW's that could not be assessed for liquid levels, similar to the 2021 semiannual monitoring events, debris was encountered at elevations that were higher than the EVGW base, thus

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precluding access to deeper portions of the EVGW's. The obstructing debris could not be identified in the field. It is noted that in 2018 these same EVGWs were assessed to be dry to each of their respective bottom elevations (approximately 1.0 to 1.5 m above the liner floor). Also, up to the 2021 semi-annual monitoring events, EVGW EV268 was determined to be dry to its depth extent, which is within 1 m above the liner floor.

### 4.1.3 Leachate Management

#### 4.1.3.1 Leachate Volume

#### **Leachate Generation:**

The primary leachate source is from precipitation infiltrating into and percolating through the waste. The groundwater contribution is negligible. Between January 1 and December 31, 2022, a total of 54,626.76 m³ of leachate was managed, of which 47,454.19 m³ of leachate was removed and transported off-Site for treatment and disposal at the Chatham Water Pollution Control Plant, while 7,172.57 m³ was irrigated onto the Poplar system during the 2022 growing season. Details regarding the 2022 Poplar System irrigation activities are discussed in **Volume 3** of the 2022 Annual Report.

A breakdown of the leachate volumes treated in 2022 is presented below.

Area of Leachate Extraction	Treatment (m³)		
Off-Site Treatme	Off-Site Treatment Management		
Existing Site MHs	613.03		
Pumping Station PS10 - From Expansion Site	41,137.26		
Pumping Station PS10 - From Existing Site	5,703.90		
Sub-Total	47,454.19		
On-Site Treatme	nt Management		
Poplar System – From Expansion Site	357.13		
Poplar System - From Existing Site	6,815.44		
Sub-Total	7,172.57		
TOTAL	54,626.76		

Note:

- 1) m³ denotes cubic metre. 1 m³ is equivalent to 1,000 L (litres).
- 2) Existing site denotes leachate extracted by tanker truck directly from one of, or a combination of, Cell 3S, 4, 5, 6, 7, 8, 9, 10/12, 11, South Cell, &/or West Cell (Sump).
- 3) The source cell leachate make-up in the Equalizations Tank, as sampled from PS10, changes over time, as outlined in Section 2.2.

Of the 54,626.76 m³ noted above, the approximate breakdown of leachate source location between the Existing Site and the Expansion Site is 24% (13,132.37 m³) and 76% (41,494.39 m³), respectively. This breakdown is based on the leachate source distribution which incorporates the separate approximate volumes of leachate extracted from the Existing Site (38% of area) and Expansion Site (62% of area) for either off-site disposal of stronger strength leachate (i.e. Expansion Site leachate) or storage of weaker strength leachate (i.e. Existing Site leachate) for use as irrigation liquid to the Poplar System during the 2022 growing season.

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### 4.1.4 Leachate Seeps & Stains

Leachate seeps or potentially ponded water within waste that were noted by WM during daily inspections were generally limited to areas adjacent to the active landfilling working area (e.g. waste mound, haul roads) of the Expansion Site in 2022. Although select observed occurrences may have been ponded water within waste, WM treated them as seeps and they were generally repaired the same day. Generally, no significant leachate seeps or stains were noted by the MECP, RWDI, or WM for the Expansion Site throughout the 2022 monitoring period. There was one (1) exception where the MECP noted a leachate seep on the south sideslope of the Expansion Site during their inspection on October 6, 2022, of which WM repaired that same day. As discussed, no leachate seeps were observed for the Existing Site during 2022.

### 4.2 Secondary Drainage Layer Hydraulic Containment

Liquid levels were recorded monthly for the SDL of Cell 1, Cell 2, Cell 4, and Cell 6A in 2022. Liquid levels for the SDL of Cell 6A were recorded beginning in September once PS8 became operational. Liquid elevations from the SDL are provided in **Table F-7** and graphically represented on **Figure F-25**, **Appendix F**.

#### **Hydraulic Containment Assessment:**

Water levels within the SDL of Cell 1, Cell 2, and Cell 4 continued to be below surrounding groundwater levels and pressures and therefore, groundwater was induced to flow toward the SDL of Cell 1, Cell 2, and Cell 4. Water levels within the SDL of Cell 6, were below surrounding groundwater levels since PS8 became operational. As the water elevations in the SDL for Cell 1, Cell 2, Cell 4, and Cell 6A, (PS2, PS4, PS6, and PS8, respectively) are typically greater than the leachate elevations within the PDL of Cell 1, Cell 2, Cell 4, and Cell 6A, the leachate in the PDLs are hydraulically contained within the PDLs from the SDLs. It is noted that in June of 2022 water taking activities for Cell 6A recompacted clayey soil liner conditioning/construction activities occurred from the SDL of Cell 4 which resulted in temporary SDL water elevation decreases. The elevations within the SDL of Cell 4 (PS6) then increased throughout Q3 and Q4 as the SDL water was not used for construction purposes during that time period. During the short period of Q3 when PS8 was operational, the water elevation in the SDL was temporarily lower than the leachate level within the PDL of Cell 6A. However, it is noted that the leachate in Cell 6A is hydraulically contained by the landfill liner which has an effective moisture content that would equate to hydraulic containment. Furthermore, the elevation within the SDL will continue to gradually increase over time as the SDL becomes saturated.

#### **Historical SDL Elevation Trends of Note:**

The initial slow increase over time in the SDL elevation within Cell 1 since its installation is a result of the large pore volume for water to accumulate within the 0.3 m thick granular layer that comprises the SDL within the western half of Cell 1. The rapid increases in SDL elevation within Cell 2, Cell 4, and Cell 6A are reflective of the very small pore volume within the 0.01 m thick Geonet that is installed as the SDL for Cell 2, Cell 4, and Cell 6A (and the eastern portion of Cell 1). It is noted that the historical periodic short duration decreasing pattern in the SDL for Cell 1, Cell 2, and Cell 4 represents water taking activities for cell-to-cell connection activities and/or soil

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moisture conditioning of the clayey soil liner material used in the construction of the cell liner system. These periodic SDL level decreases are expected during the Expansion Site construction and as discussed, do not represent a concern as once a cell is excavated, groundwater movement is toward the cell excavation, thereby hydraulically containing any liquid (both in SDL and PDL) in the cell.

### 4.3 Active Aquitard Groundwater Movement

The active aquitard represents a shallow groundwater flow system whereby precipitation infiltrates into soil fractures, root casts, and other weathering features and moves in a dominantly lateral direction. The low permeability of the surrounding silty clay to clayey silt matrix reduces the volume of groundwater movement through the active aquitard. The low volume of water within the active aquitard is also evident by the slow recovery of monitoring wells after purging and sampling.

Groundwater elevations measured during the 2022 annual monitoring program are presented with historical data in **Table F-4**, **Appendix F**.

#### 4.3.1 Groundwater Elevation Patterns

Groundwater elevations measured for the Active Aquitard on May 2, 2022, are presented on **Figure 4**. Considering the north-south flow divide through the Poplar Plantation and historical patterns of groundwater movement, an easterly and westerly direction of groundwater movement is inferred. Surficial features such as ditches, waste cells, excavations, ponds, berms, and the leachate management systems, as well as precipitation amounts, will locally influence the groundwater flow direction.

#### 4.3.2 Groundwater Elevation Trends

Groundwater elevation hydrographs for the active aquitard are presented on **Figures F-13** through **F-18**, **Appendix F**. Long-term trends are summarized in **Table 4**. For the purpose of assessing elevation trends, a constant elevation trend is defined as having 0.1 m or less seasonal variation.

In the long-term within the active aquitard, groundwater elevations have continued to fluctuate. Overall, fluctuating patterns in groundwater elevation were noted across the Site in the Active Aquitard, which are attributed to seasonal effects from precipitation and snowmelt. Overall, groundwater elevations were generally noted to be within their historical ranges. The groundwater elevation at OW72-6 was noted to have increased to a new high in November 2022.

### 4.4 Interstadial Silt & Sand Groundwater Movement

The interstadial silt and sand hydrostratigraphic unit consists of silt and sand layers or lenses within the interstadial deposit that overlies the Rannoch Till. Groundwater movement within the interstadial silt and sand will be controlled by hydraulic gradients and the soil texture. For the purpose of the landfill compliance assessment, it is assumed that the silt and sand layers or lenses are hydraulically connected and that groundwater movement is in the direction of the inferred hydraulic gradient.

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Groundwater elevations measured during the 2022 annual monitoring program are presented with historical data in **Table F-4**, **Appendix F**. Hydrographs for the interstadial silt and sand are provided on **Figures F-19** through **F-22**, **Appendix F**.

#### 4.4.1 Groundwater Elevation Patterns

Groundwater elevations measured for the Interstadial Silt and Sand on May 2, 2022 are presented on **Figure 5**. Groundwater pressures suggest a consistent pattern to the historical interpretation of an easterly and westerly groundwater flow direction from a north-south groundwater divide that extends from Confederation Line to Zion Line (Jagger Hims Limited, 2005). However, the excavation of Cell 1, Cell 2, Cell 4, and Cell 6A has reduced the groundwater potentiometric pressures within the interstadial silt in these areas and locally induces groundwater to move toward and into the SDL of the cells. These lowered groundwater pressures are expected to return to pre-existing conditions as the soil mass is replaced with waste mass. Higher groundwater pressure elevations relative to other areas may occur below the Central Fill Area (defined as waste cells: Cell 3S, Cell 3 through Cell 7, as well as Cell 9 and Cell 11) of the Existing Site due to the mass of the overlying waste and leachate, thereby hydraulic gradients are expected to be upward toward the waste.

#### 4.4.2 Groundwater Elevation Trends

Long-term elevation trends are summarized in **Table 4**. For the purpose of assessing elevation trends, a constant elevation trend is defined as having 0.1 m or less seasonal variation.

Groundwater elevations within the interstadial silt and sand have generally fluctuated in the long-term, with the elevations at OW54-10, OW67-11, OW72-10, and OW73-9 exhibiting generally decreasing and fluctuating trends over time. It is noted that the groundwater elevations at the aforementioned monitoring wells appear to be stabilizing. Ongoing monitoring will evaluate the noted trends over time.

For the monitoring wells where seasonal fluctuations in the groundwater elevation have historically been observed, this trend generally continued in 2022. For these wells, the November 2022 groundwater elevations within the interstadial silt and sand were generally lower or similar compared to those observed in May 2022. Within the interstadial silt and sand flow system, the overlying waste and leachate levels, as well as the Cell 1, Cell 2, Cell 4, and Cell 6A excavations, induced localized potentiometric pressure increases and decreases, which also resulted in variations in the local flow direction. Overall, with the exception of where decreasing trends continued to be observed in 2022, groundwater elevations were generally noted to be within their historical ranges.

### 4.4.3 Vertical Hydraulic Gradients

Vertical hydraulic gradients below the Site in the active aquitard to the interstadial silt and sand hydrostratigraphic unit were calculated for the May 2 and November 1, 2022, data.

The hydraulic gradients that were calculated are presented in **Table F-5**, **Appendix F**. In summary, vertical hydraulic gradients continued to be dominantly downward toward the interface aquifer around the waste footprints and below the Existing Site. However, localized upward hydraulic gradients occurred from the active

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aquitard to the interstadial silt and sand at the monitoring well locations summarized in the table below. Ongoing monitoring will continue to evaluate these trends over time.

May 2022	November 2022		
OW80-3/OW80-6	OW40D-4/OW40A-7 OW84-6/OW84-11		
	OVV04-0/OVV04-11		

### 4.5 Interface Aquifer Groundwater Movement

Occasional discontinuous layers of sand and gravel between the Rannoch Till and underlying bedrock constitute the basal sand. For the purpose of the landfill compliance assessment, it is assumed that groundwater movement is in the direction of the inferred hydraulic gradient. The upper fractured portion of the bedrock and the overlying basal sand form the local bedrock aquifer and are generally referred to as the interface aquifer.

Groundwater elevations measured during the 2022 annual monitoring program are presented with historical data in **Table F-4**, **Appendix F**. Hydrographs are provided on **Figures F-23** to **F-24**, **Appendix F**.

#### 4.5.1 Groundwater Elevation Patterns

Groundwater elevations measured for the Interface Aquifer on May 2, 2022, are presented on **Figure 6**. Overall, a southwesterly groundwater flow direction was apparent below the Existing Site and the Expansion Site. This flow direction is consistent with the historical inferred direction of groundwater movement within the interface aquifer (Jagger Hims Limited, 2005). Of note, as more overburden soil is removed during landfill cell construction, interface aquifer pressures are expected to decrease temporarily. As more waste is deposited within the waste cells of the Expansion Site, the increase in overlying mass will increase the groundwater pressures of the bedrock aquifer.

#### 4.5.2 Groundwater Elevation Trends

Long-term elevation trends are summarized in **Table 4**. For the purpose of assessing elevation trends, a constant elevation trend is defined as having 0.1 m or less seasonal variation. Groundwater pressure elevation trends and patterns for the interface aquifer continue to be generally consistent with historical findings with a fluctuating trend with time. Additionally, the majority of monitoring wells in the interface aquifer were showing a fluctuating and decreasing trend from about 2010 to 2018. This pattern at these locations was expected as it indicates a response to depressurization from the removal of overlying soil within the Expansion Site waste footprint. It is anticipated that these decreasing elevation trends will be temporary, and pressures will begin to increase once a sufficient waste mass is landfilled to replace the excavated clayey soil mass. As expected, monitoring wells OW39A-26, OW79-26, OW80-27, and OW81-27 are showing a fluctuating and increasing trend since about 2018 as a significant waste mass has been progressively landfilled into the southern half of the Expansion Site. Overall, with the exception of where a more dominant fluctuating trend continued to be observed in 2022, groundwater liquid levels were generally noted to be within their historical ranges.

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### 4.5.3 Vertical Hydraulic Gradients

Vertical hydraulic gradients below the Site in the interstadial silt and sand to the interface aquifer hydrostratigraphic unit were calculated for the May 2 and November 1, 2022 data. The hydraulic gradients that were calculated are presented in **Table F-5**, **Appendix F**. In summary, vertical hydraulic gradients were noted to be downward during both the May and November 2022 monitoring events from the Interstadial Silt and Sand layer to the Interface Aquifer owing to the low leachate and groundwater levels within Cells 1, 2, 4 and 6A. Ongoing monitoring will continue to evaluate these trends over time.

## 5 CHEMICAL & GAS MONITORING RESULTS

### 5.1 Leachate

In accordance with the landfill EMP, leachate sampling from within select maintenance holes across the Existing Site was completed on May 6, 2022, as part of the annual monitoring requirement. A leachate sample was obtained from the Sump that is located within the West Cell. A second leachate sample designated CFA-Comp consisted of the collection of a composite sample obtained from MH3SA-F, MH4B, MH5A, MH6A, MH7A, MH9A, and MH11A to represent leachate within the Central Fill Area. Leachate is also sampled from MH18 to represent leachate from the South Cell.

Samples were also collected from PS1, PS3, and PS5 during the spring monitoring event on May 10, 2022. PS1, PS3, and PS5 samples were collected directly from the pumping station sump. Chemical results are summarized in **Appendix G**.

**Table 6** provides a chemical summary of the long-term leachate characteristics for the Existing Site and the Expansion Site to November 2022. The 2022 leachate chemical results were generally within the respective historical ranges for the parameters analyzed.

As shown on **Figure G-1**, **Appendix G**, the chloride concentrations in leachate from the Equalization Tank show a fluctuating trend since 2013. Concentration fluctuations over time are expected due to the nature of leachate, as well as to the variations in the relative contribution of weaker leachate from new waste (i.e., new waste in Cell 6A) or from the monofill cells (i.e., Cells 10/12) of the Existing Site to stronger leachate from the aging waste (i.e., waste in Cell 1) or from the West Cell (Sump) of the Existing Site. Chloride concentrations in leachate from the Equalization Tank will continue to be monitored during future sampling events.

#### 5.1.1 Leachate Indicator List

In general, leachate concentrations vary across the Site, with stronger leachate within Cell 6 and the West Cell compared to other cells of the Existing Site. Leachate constituent concentrations for the Expansion Site are generally within the low range for concentrations detected in the Existing Site. This difference is attributed to the relatively young age of the waste (compared to waste in the Existing Site) and the onset of waste decomposition. The 2022 and historical ranges in chemical concentrations for the leachate, as well as the background

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groundwater and surface water, are presented in **Table 6**. Based on a comparison of the leachate concentrations to the background groundwater and surface water concentrations, the following parameters have had notably elevated concentrations within the leachate.

- EC
- Alkalinity
- Calcium
- Magnesium
- Sodium
- Potassium
- Chloride
- Dissolved Organic Carbon
- Ammonia
- Total Kjeldahl Nitrogen
- Boron
- BTEX Compounds (Benzene, Toluene, Ethylbenzene, and Xylenes)

A landfill leachate effect on groundwater or surface water quality is defined as leachate mixing with the water and migrating in the direction of water movement beyond the Site boundary. To differentiate landfill leachate effects from natural water quality variability, a group of parameters indicative of landfill leachate was selected based on the following items.

- Elevated concentrations in leachate.
- Mobility and stability in groundwater and surface water.
- Representative of a larger chemical group (e.g., boron for metals).
- Applicable Ontario Drinking Water Standard (ODWS) or PWQO.

A group of indicator parameters selected for this landfill, termed the Primary Leachate Indicator List (PLIL), are summarized below.

Primary Leachate Indicator List (PLIL)				
PLIL – Groundwater	PLIL – Surface Water			
Chloride Nitrate (as N) Boron	Chloride Ammonia (unionized) Phenols Boron Nickel Chromium (total) Zinc			

A supplemental group of parameters, termed the Secondary Leachate Indicator List (SLIL), is selected based on elevated concentrations within the leachate. The SLIL is used for quality assurance of the chemical database, for input to establish trigger concentrations, and for further data analyses in the event of trigger concentration exceedances. The SLIL parameters are summarized below.

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	Secondary Leachate Indicator List (SLIL)						
SLIL - Gr	SLIL – Groundwater SLIL – Surf						
Alkalinity Sulphate Calcium Magnesium Potassium Sodium Barium Iron DOC	Ammonia (total) TKN pH Conductivity Cadmium Lead Benzene Toluene Ethylbenzene	Alkalinity Sulphate Calcium Magnesium Potassium Sodium Iron Nitrate TKN	TDS pH Conductivity Turbidity Dissolved Oxygen				
TDS	Xylenes	Phosphorus (total)					

The concentrations of leachate constituents will vary with time. Leachate quality monitoring will continue at the Existing Site and the Expansion Site to assess the changing leachate characteristics for a landfill impact assessment and for leachate management.

#### 5.2 Groundwater

Field groundwater chemical results for the May, November, and supplemental monitoring events are presented in **Tables H-1** to **H-3**, **Appendix H**. The 2022 laboratory chemical results are presented in **Tables H-4** and **H-5**, **Appendix H**. Historical data are provided for comparison. Laboratory reports of analysis and QA/QC documentation are included as **Table H-6**, **Appendix H**.

Groundwater quality monitoring occurred at the required monitoring locations in consideration of the groundwater monitoring schedule within the approved landfill EMP. An updated monitoring schedule is presented within **Table B-2**, **Appendix B**, which shows the monitoring wells required for sampling, the sampling frequency, and the parameters for analytical testing. Monitoring well designations and locations are presented on **Figure 2**. Monitoring well nest OW81 was installed in June 2019; with monitoring of groundwater quality beginning in May 2020. Monitoring well nests OW82, OW83, and OW84 were installed in June 2022; with monitoring of groundwater quality scheduled to begin in May 2023.

#### 5.2.1 General Chemical Trends

To assess long-term chemical trends with time, time-concentration graphs for chloride, nitrate, and boron were prepared. Based on the background concentrations presented in **Table 6**, concentrations were considered constant where results varied by less than 15 milligrams per litre (mg/L) for chloride, 0.5 mg/L for nitrate, and 0.2 mg/L for boron. Concentration trends that varied by greater than these concentration ranges were interpreted as increasing, decreasing, or fluctuating.

Concentration trends for the time-concentration graphs on **Figures H-1** to **H-14**, **Appendix H**, are summarized in **Table 7**. Typically, the long-term concentration trends indicate constant or fluctuating concentrations with time, with some exceptions, which are discussed below. It is noted that the below observations do not represent immediate concerns, but as a matter of diligence will be evaluated over time to verify that a trend of concern does not develop or corrective measures, if required can be implemented.

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#### **Active Aquitard:**

- At monitoring well OW16-6, the chloride concentrations have been generally stable after the short-term historical high concentration of 77 mg/L noted in May 2015, which subsequently decreased to an approximate average of 45 mg/L since then. It is noted that the chloride concentrations discussed are notably less than the trigger concentration of 106 mg/L.
- At monitoring well OW17-4, the chloride concentrations have been generally stable after the short-term historical high concentration of 71 mg/L noted in May 2016, which subsequently decreased to an approximate average of 30 mg/L since then. It is noted that the chloride concentrations discussed are notably less than the trigger concentration of 106 mg/L.
- At monitoring well OW54A-4, chloride concentrations have been increasing and fluctuating since late 2014. It is noted that the aforementioned chloride concentrations have shown an approximate stabilization at 26 mg/L since 2018 and are notably less than the trigger concentration of 106 mg/L.
- At monitoring well OW56-4, nitrate concentrations have generally been low and constant, however, isolated elevated concentrations have been detected over time; with the most recent event occurring in 2018. It is likely that the nitrate concentrations are a result of effects resulting from fertilizer application to the adjacent field. Nitrate concentrations have continued to be low and constant since 2018.
- At monitoring well OW58-6, nitrate concentrations have been low and constant, however, an isolated event with a concentration that was a new historic high was detected in May 2022. It is noted that the aforementioned nitrate concentration of 0.47 mg/L is well below the trigger concentration of 2.3 mg/L. It is likely that the nitrate concentration detected in May 2022, was a result of effects resulting from fertilizer application to the adjacent field. The May 2022 nitrate concentration is not a concern as the detected November 2022 concentration was low, as normally observed.
- At monitoring well OW67-4, the concentrations of chloride, nitrate, and boron have exhibited generally
  decreasing and fluctuating trends over time. Periodic increases in concentrations for each parameter
  have occurred over time, including chloride in November 2022. However, the overall trend for each
  parameter has generally been decreasing and fluctuating since monitoring began at OW67-4. It is noted
  that the concentrations of chloride, nitrate, and boron were below their respective trigger concentrations
  in 2022.
- At monitoring well OW69-5, boron concentrations continue to show a generally constant but slightly fluctuating trend with concentrations hovering around the shallow groundwater trigger concentration of 1.1 mg/L. At this time the boron concentrations at OW69-5 are not a concern as it is expected the more mobile parameter chloride would be observed first if the source of the boron were leachate. It is likely that, similar to observations for monitoring well OW58-14, that was decommissioned in 2016, the monitoring well's bentonite seal is moving into the filter screen material of the monitoring well, and as such may require decommissioning and replacement in the future depending on chemical results.
- At monitoring well OW71A-5, the chloride concentrations are showing an increasing trend to a historical high of 32 mg/L in November 2022. However, the chloride concentrations are notably less than the Active Aquitard trigger concentration of 106 mg/L. It is noted that monitoring well OW71A-5 is not a compliance monitoring well for the Site but was added to supplement historically dry conditions at

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OW67-4. The recent chloride concentrations at OW71A-5 and OW67-4 are generally similar and the concentrations of the other PLIL parameters are not showing a similar trend.

#### **Interstadial Silt and Sand:**

- At monitoring well OW46-7, the concentration of boron has been observed to have been fluctuating and increasing since November 2019, which is similar to the short-term increase in 2015. However, the boron concentrations are notably less than the historical concentrations prior to 2003. The boron concentrations remain below the trigger concentration of 2.1 mg/L. The other PLIL parameters are not elevated or showing an increasing trend. It is likely that the observed periodic increases in boron concentrations is an effect from the bentonite seal moving into the filter pack of the monitoring well. At historical monitoring well OW58-14, the observed boron concentrations were proven to be related to bentonite seal impacts from the monitoring well seal moving into the sand filter pack. Consequently, per MECP approval, OW58-14 was decommissioned and replaced as OW58-17, which has shown constant and acceptable concentrations of boron since its installation in 2014.
- At monitoring well OW47-6, the concentrations of chloride have been constant and low since data has been recorded. During the May 2022 monitoring event, a new historical maximum was recorded at monitoring well OW47-6. It is noted that the aforementioned chloride concentration (14 mg/L) is well below the trigger concentration of 116 mg/L.
- At monitoring well OW67-11, chloride concentrations showed an increasing trend between about 2009
  and 2013 and have since showed a generally fluctuating trend. Nitrate concentrations have continued to
  fluctuate since monitoring began. Boron concentrations have fluctuated and decreased since about
  2011. The chloride concentrations are notably below the trigger concentration of 116 mg/L for the
  interstadial silt and sand hydrostratigraphic unit.
- At monitoring well OW80-6, the concentrations of chloride, nitrate, and boron detected in May 2022 were
  noted to be anomalous due to overall groundwater quality showing significant deviations (typically
  lesser) from historical concentrations. Further monitoring will evaluate these parameter concentrations
  over time.
- At monitoring well OW81-7, chloride concentrations are elevated as a result of effects road salting along Nauvoo Road. It is noted that due to road salt influences at this location, chloride was removed as part of Trigger Assessment per MECP letter, dated August 18, 2022.

### 5.2.2 Organic Chemistry

The 2022 and historical organic chemical results are provided in **Table H-5**, **Appendix H**. Tested organic chemical constituents at each monitoring location were not detected at concentrations greater than their respective laboratory RDL during 2022.

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### 5.2.3 Trigger Mechanisms

Groundwater compliance is assessed based on criteria calculated with respect to the Ministry of the Environment and Energy's (MOEE) Guideline B-7 Reasonable Use Concept (Guideline B-7) and evaluated at the Site boundaries (Points of Compliance). Trigger concentrations are used to assess potential sources of concentration changes. The groundwater trigger concentrations are based on 80% of the calculated Guideline B-7 criteria for the reasonable use of groundwater (MOE, 1994), in accordance with Condition 14.2 of the ECA. When background groundwater quality satisfies the ODWS, background groundwater quality cannot be degraded by more than 25% of the ODWS for health-related parameters (e.g., nitrate) or by more than 50% of the ODWS for aesthetic-related parameters (e.g., sodium). Where background concentrations naturally exceed the ODWS, background groundwater quality becomes the criteria.

Historical background groundwater quality for local monitoring wells was used to calculate the trigger concentrations at the Site. As a screening process, trigger concentrations for the PLIL were calculated and assessed as provided in Table 3 of the landfill EMP.

#### 5.2.3.1 Points of Compliance

An assessment of potential landfill leachate effects on groundwater quality is completed at the Site boundary to encompass the Expansion Site footprint and operations, as well as to consider potential effects of the Existing Site on groundwater quality. Monitoring wells are used at strategic locations around the Site to obtain an acceptable representation of groundwater quality moving beyond the Site boundary. These strategic locations are called the Points of Compliance and are assessed in accordance with Guideline B-7.

To account for the natural groundwater quality variability, the trigger concentrations consider the PLIL parameters discussed in **Section 5.1.1**. Exceedance of the trigger concentrations at the monitoring wells initiates the assessment process, which is detailed within Figure 3 of the landfill EMP. **Tables 8** to **10** provide the groundwater trigger concentration comparison and the 2022 chemical concentrations for the PLIL parameters.

Verification groundwater monitoring at the Site adheres to Condition 14.4 of the Waste ECA, which references the landfill EMP. Groundwater verification is required to be completed for two (2) subsequent events separated by six months with the first verification monitoring event taking place within one month of the initial sampling event.

#### 5.2.3.2 Trigger Concentration Assessment

#### **Primary Leachate Indicator List Comparison:**

Based on the chemical analytical results for the 2022 monitoring events, the concentrations of the PLIL parameters at the groundwater monitoring wells, including the Points of Compliance, within the active aquitard, the interstadial silt and sand, as well as the interface aquifer satisfied the relevant trigger concentrations, with one (1) exception at OW81-7.

At monitoring well OW81-7, a concentration of chloride was detected that was greater than the respective trigger concentration during the spring monitoring event.

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Shallow groundwater quality at the location of monitoring nests OW80 and OW81 are interpreted to be impacted by nearby road salting activities on Nauvoo Road. The MECP approved the discontinuation of utilizing the parameter chloride as a trigger mechanism OW80-3 and OW80-6 in 2010 following their installation in 2009. Given its similar construction and proximity to Nauvoo Road, the groundwater at interstadial silt and sand monitoring well OW81-7 is interpreted to also be influenced by road salting activities. As such, the chloride concentration of 200 mg/L in this monitoring well that was detected in May 2022 is not attributed to landfill leachate related impacts. This is also evidenced by the fact that the liquid level elevation noted at monitoring well OW81-7 was higher in elevation (235.15 mASL) than the secondary drainage layer (SDL) at pumping stations PS4 (229.94 mASL) and PS6 (227.73 mASL) in May 2022, indicating groundwater flow toward the landfill.

As noted in the 2022 Spring Groundwater Quality Monitoring Letter of Notification, prepared by RWDI and submitted to the MECP on June 3, 2022, a groundwater trigger concentration exceedance for chloride was identified within the groundwater monitoring well OW81-7. Based on this noted chloride groundwater trigger concentration exceedance, per the process outlined in the Environmental Monitoring Program (Jagger Hims Limited, 2007) (EMP), a verification monitoring event for the primary and secondary leachate indicator list parameters was required per Figure 3 of the EMP. Verification monitoring on June 2, 2022, confirmed the initial assessment findings that the groundwater quality at OW81-7 is being impacted by road salting effects along Nauvoo Road. As discussed above, based on the road salt influences in this area, chloride is no longer required to be evaluated as a PLIL parameter, per MECP approval on August 18, 2022. As such a second verification sample in November was not required to be collected.

#### Secondary Leachate Indicator List Assessment:

Based on the chemical analytical results for the 2022 monitoring events, the concentrations of the SLIL parameters at the groundwater monitoring wells, including the Points of Compliance, within the active aquitard, the interstadial silt and sand, as well as the interface aquifer satisfied the relevant trigger concentrations, with one (1) exception at OW80-6 and two (2) exceptions at the 'Cemetery Well'.

With respect to the elevated lead concentration noted within the groundwater at monitoring well OW60-4 the past seven (7) consecutive sampling events, the MECP approved to discontinue the use of lead as a SLIL parameter for groundwater at the location of OW60-4 per a letter dated October 2, 2020. Therefore, similar to chloride removal from the trigger concentration assessment process at a few monitoring wells, lead was removed from the trigger concentration assessment at OW60-4 and the detected values will be documented for tracking purposes.

The detected cadmium concentration at monitoring well OW80-6 is not a concern as cadmium is a SLIL parameter, and the PLIL parameters (chloride, boron, and ammonia) were not detected at elevated concentrations. In addition, given that the majority of the chemical parameter concentrations deviated significantly from their historical trends with the majority of parameter concentrations decreasing, it expected that the results are anomalous. As such, the cadmium exceedance in the groundwater at monitoring well OW80-6 is interpreted to not be landfill leachate related. This is also evidenced by the fact that the liquid level elevation noted at monitoring well OW80-6 was higher in elevation (235.44 metres above sea level (mASL) and) than the

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secondary drainage layer (SDL) at pumping stations PS4 (228.07 mASL) and PS6 (227.73 mASL). As such, groundwater was noted to be induced, by design, toward the landfill.

The detected cadmium and lead concentrations at the 'Cemetery Well' monitoring location are not a concern as cadmium and lead are SLIL parameters, and the PLIL parameters (chloride, boron, and ammonia) were not detected at elevated concentrations. As such, the cadmium and lead exceedances in the groundwater at the 'Cemetery Well' monitoring location are interpreted to not be landfill leachate related.

Of note, there was a significant amount of precipitation (17.6 mm) that occurred at the Site on May 3, 2022, prior to the samples being collected at monitoring well OW80-6 and at the 'Cemetery Well' monitoring location on May 5 and May 6, 2022, respectively. It is possible that the significant amount of precipitation had an influence on the groundwater quality sampled at these locations. As cadmium and lead are SLIL parameters, verification monitoring is not required at monitoring well OW80-6 and at the 'Cemetery Well' monitoring location. The detection of cadmium and lead at these monitoring wells is noteworthy and will warrant observation of focus during the next sampling event. As the 'Cemetery Well' showed effects similar to OW80-6, which is developed in the Interstadial Silt and Sand, it is possible that the 'Cemetery Well' is also installed within this hydrostratigraphic unit and is not screened within the bedrock aquifer. Caution will be used when assessing groundwater quality at the 'Cemetery Well'

During the 2022 monitoring events, VOCs tested within the groundwater at the monitoring well locations satisfied their respective trigger concentrations.

In summary, based on the chemical analytical results for the 2022 monitoring events, the concentrations of the PLIL and SLIL parameters at the groundwater monitoring wells, including the Points of Compliance, within the active aquitard, the interstadial silt and sand, as well as the interface aquifer generally satisfied the relevant trigger concentrations.

#### 5.3 Surface Water

Surface water chemical results are summarized in **Tables I-1** to **I-4**, **Appendix I**. Laboratory reports and documentation of quality control are provided in Table I-5, Appendix I.

#### 5.3.1 General Chemical Trends

As discussed in Section 5.1.1, there are seven (7) surface water PLIL parameters that represent effective indicators of landfill leachate effects on surface water quality for the Site. To assess chemical trends with time, time-concentration graphs for chloride, un-ionized ammonia, boron, and zinc were prepared.

Data were analyzed for long-term trends, which considered the historical data. Concentrations were considered constant where results varied by less than 15 mg/L for chloride, 0.02 mg/L for un-ionized ammonia, 0.2 mg/L for boron, and 0.02 mg/L for zinc. Concentrations trends that varied by greater than these concentration ranges were interpreted as increasing, decreasing, or fluctuating. A minimum of three (3) to five (5) data points are required to determine a chemical trend with time. Concentration trends are summarized in Table 11. Timeconcentration graphs are presented on Figures I-1 to I-4, Appendix I.

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Long-term concentration trends indicate both constant and fluctuating concentrations with time. Fluctuating concentrations generally showed a seasonal pattern for boron and chloride concentrations at stations SS1 and SP1. Chloride concentrations at SP2 and SP4 appear to have a fluctuating pattern in the long-term. Additionally, in the long-term, unionized ammonia concentrations continued a constant trend of being below their respective laboratory RDL and/or less than the respective PWQO (0.02 mg/L), with occasional historical exceedances at SS1, SP1, and SP2.

Concentrations of chloride, boron, unionized ammonia, and zinc in surface water within the Sedimentation Ponds and at compliance point SS1 were typically similar to each other, with some recent exceptions where elevated and fluctuating concentrations of boron have generally been noted at SS1, SP1, and SP2. The noted boron concentrations in 2022 at SS1, SP1 and SP2 are not a concern as they are not landfill leachate related but are a result of short-term effects from general erosional effects from significant rainfall, erosion effects related to Street C construction in 2021 that was upstream of SS1 and SP2, and sediment removal activities from Sedimentation Ponds 1 and 2 in the winter of 2021 and the late fall of 2022. The noted unionized ammonia concentration in Q4 2022 at SP2 is also not a concern as it is not landfill leachate related but is a result of short-term effects of the de-watering and sediment removal activities that occurred in Sedimentation Pond 2 in the late fall of 2022. Further details pertaining to individual surface water monitoring events conducted in 2022 are provided in **Section 5.3.5**.

### 5.3.2 Organic Chemistry

Organic chemical results are provided in **Table I-3**, **Appendix I**. Tested organic chemical constituents at the downstream compliance monitoring stations SS1, SP2, SP3, and SP4, as well as the internal compliance monitoring station SP1, were not detected at concentrations greater than their respective laboratory RDL.

### 5.3.3 Biomonitoring Program - Annual Spring Routine Event

The laboratory results for the biomonitoring monitoring events completed in 2022 are provided within **Table I-6**, **Appendix I**. The sample aliquots collected to assess biological conditions were assessed for surface water quality toxicity to *daphnia magna* and rainbow trout. The biological testing results showed that there was not a potential for detrimental effects to aquatic life in the discharge water.

### 5.3.4 Biomonitoring Program - Verification Events

Biomonitoring was also conducted in 2022 in accordance with conditions approved in the 2014 MECP Letter. Surface water verification sampling included the collection of a grab sample for analysis of the toxicity to rainbow trout and *daphnia magna* associated with verification surface water monitoring events for compliance monitoring discussed in **Section 2.4.3**.

During the verification sampling events, the verification biological results indicated that the surface water was of acceptable quality for continued discharge.

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### 5.3.5 Trigger Concentration Assessment

**Table 12** provides the Post 2021 trigger concentrations utilized to assess surface water quality during 2022, as well as the Post 2022 trigger concentrations that will be utilized in 2023 to assess the chemical quality of surface water.

A comparison of the 2022 surface water quality with the Post 2022 trigger concentrations is provided in **Table I-4**, **Appendix I**. Outlined below is a detailed discussion of the trigger concentration assessments for each surface water monitoring event, by calendar quarter for 2022. A quality comparison for background stations SS10 and SS16 is also provided where applicable.

#### First Quarter - January 1 to March 31, 2022

#### February 17, 2022:

Surface water samples were collected at compliance monitoring stations SS1, SP2, SP3, and SP4, as well as internal assessment monitoring location SP1 and background monitoring station SS16. Surface water samples were not collected at background monitoring station SS10 due to there not being flowing conditions to initiate sample collection. The surface water quality at the required monitoring stations generally satisfied the relevant trigger concentrations, with one (1) exception. For the surface water sample collected at SS1, the concentrations of nickel (0.039 mg/L), total chromium (0.026 mg/L), and zinc (0.07 mg/L) were greater than their respective trigger concentrations (0.027 mg/L, 0.024 mg/L, and 0.06 mg/L, respectively). As such, verification monitoring at compliance monitoring stations SS1 was required. The verification monitoring was completed for SS1 on May 4, 2022, and is discussed below.

The elevated nickel, chromium, and zinc concentrations noted at the location of surface water monitoring station SS1 were interpreted to be as a result of erosional effects, in part due to construction activities related to the resurfacing of the street directly west and south of Cell 1 and Cell 2 of the Expansion Site (Street C), which included the re-grading of the street-side side slopes of the surface water drainage ditches along Street C. These construction works were completed in the fall of 2021 and would result in temporary increased erosional effects as the work required the removal of much of the established vegetation and straw-bale check dams that would normally provide erosion control in this area. As the vegetation was in the early stages of re-growth, it was expected that the temporary increased erosional effects could still be observed in the near term following Q1. As seen in **Figure 3**, the surface water drainage ditch on the south side of Street C leads directly to surface water monitoring station SS1. In addition, of the significant amount of precipitation (20.8 mm) that occurred prior to sampling, it is noted that the majority of the precipitation fell within 6 hours of sample collection and precipitation was also occurring during the morning of the day of sample collection. Along with the intense rainfall, partial snowmelt was observed to be contributing to the streamflow. The visually identifiable turbid waters for the streamflow at SS1 was confirmed as such through field measurements for turbidity.

Though elevated, the laboratory and field chemical constituent concentrations were noted to be within the historical range for surface water collected at monitoring station SS1. In addition, some straw-bale check dams were re-installed upstream of monitoring station SS1 in the early winter of 2021. As a result, the surface water

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quality was observed to have generally improved. It is noted that the snow and ice build-up throughout the ditches, that was only partially melted, impacted the functionality of the straw bale check dams. Overall, it was interpreted that the elevated nickel, total chromium and zinc concentrations at monitoring station SS1 are not as a result of a landfill leachate effect, but rather from soil erosional effects.

#### May 4, 2022:

A verification surface water sample was not collected at monitoring station SS1 during the Q1 as there were no precipitation events before the end of Q1 that generated flowing conditions following the routine sampling event that occurred on February 17, 2022. To address the boron concentrations at SS1, the routine sample for monitoring station SS1 that was collected during Q2 on May 4, 2022, also was assessed as the verification sample for SS1 from Q1. The verification monitoring indicated an exceedance of boron again detected, however the biological results indicated that the surface water quality does not pose a detrimental effect to aquatic life and that the surface water was of acceptable quality for continued discharge and that no further verification monitoring was required (refer to **Section 5.3.3** and **Section 5.3.4**).

#### Second Quarter - April 1 to June 30, 2022

#### May 4, 2022:

Surface water samples were collected at compliance monitoring stations SS1, SP2, SP3, and SP4, as well as internal assessment monitoring location SP1 and background monitoring stations SS10 and SS16. The surface water quality at the required monitoring stations generally satisfied the relevant trigger concentrations, with two (2) exceptions. For the surface water samples collected at SS1 and SP2, the concentration of boron (0.36 mg/L and 0.39 mg/L, respectively) were greater than their respective trigger concentration (0.20 mg/L). As such, verification monitoring at compliance monitoring stations SS1 and SP2 was required. The verification monitoring was completed for SS1 and SP2 on May 16, 2022, and is discussed below.

An evaluation of the source of boron was completed and it was determined that the elevated boron concentrations noted at the locations of surface water monitoring stations SS1 and SP2 were interpreted to be partly attributed to erosional effects at the Site, as well as from sediment removal from Sedimentation Ponds 1 and 2 in the winter of 2022. Sediment removal from the forebays of Sedimentation Ponds 1 and 2 was completed in February 2022, as well as for a small portion of the main forebay of Sedimentation Pond 2 in March 2022. It is noted that SS1 is downstream of Sedimentation Pond 1, as shown in **Figure 3**. The disturbance of the sediment in these locations, which would have boron bound into the sediment, would have caused a loading of boron in a dissolved state to the water within these ponds, which would dimmish over time with incoming precipitation runoff.

Though elevated, the boron concentrations were noted to be within the historical range for surface water collected at monitoring stations SS1 and SP2. In addition, some straw-bale check dams were re-installed upstream of monitoring station SS1 in the early winter of 2021. As a result, the surface water quality was observed to have generally improved. Overall, it was interpreted that the elevated total boron concentrations at monitoring stations SS1 and SP2 was not as a result of a landfill leachate effect, but rather from soil erosional effects and sediment removal activities from Sedimentation Ponds 1 and 2.

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#### May 16, 2022:

To address the boron concentrations at SS1 and SP2, a surface water verification monitoring event was completed on May 16, 2022. The verification monitoring for SS1 indicated an acceptable concentration of boron whereas the verification monitoring for SP2 indicated a boron concentration greater than the trigger concentration (0.20 mg/L). However, the verification monitoring for both SS1 and SP2 denoted acceptable biological results indicating that the surface water quality does not pose a detrimental effect to aquatic life and that the surface water was of acceptable quality for continued discharge and that no further verification monitoring was required (refer to **Section 5.3.4**).

#### Third Quarter - July 1 to September 30, 2022

#### July 20, 2022:

Surface water samples were collected at compliance monitoring station SP2, as well as internal assessment monitoring location SP1 on July 20, 2022, as part of the routine monitoring program following a precipitation event of greater than 10 mm in a 24-hour period from 8:00 AM July 19 to 8:00 AM July 20. Samples were not collected at surface water monitoring stations SS1, SS10, SS16, SP3, and SP4 due to there not being flowing conditions to initiate sample collection. As such, monitoring of SS1, SS10, SS16, SP3, and SP4 remained required to be completed. The surface water quality at the required monitoring stations satisfied the relevant trigger concentrations with two (2) exceptions at compliance monitoring station SP2 and internal assessment monitoring location SP1, where the concentrations of boron (0.40 mg/L and 0.43 mg/L) were greater than the trigger concentration (0.20 mg/L). Therefore, verification monitoring was required for SP1 and SP2, as discussed below was completed on August 2, 2022.

An evaluation of the sources of boron was completed and it was determined that the elevated boron concentrations noted at the locations of surface water monitoring stations SP1 and SP2 was interpreted to partly be as a result of sediment removal from Sedimentation Ponds 1 and 2 in the winter of 2022. Sediment removal from the forebays of Sedimentation Ponds 1 and 2 was completed in February 2022, as well as for a small portion of the main forebay of Sedimentation Pond 2 in March 2022. The disturbance of the sediment in these locations, which would have boron bound into the sediment, would have caused a loading of boron in a dissolved state to the water within these ponds. Since the completion of the Q2 surface water monitoring event, that was completed in May 2022, there was a lack of runoff due to infrequent periods of precipitation. As a result, the water in the sedimentation ponds had generally been stagnant and had not received typical incoming runoff volumes to facilitate routine flushing.

Though elevated, the laboratory and field chemical constituent concentrations were noted to be within the historical ranges for surface water collected at monitoring stations SP1 and SP2. Thus, it is interpreted that the elevated boron concentrations at monitoring stations SP1 and SP2 are not a result of a landfill leachate effect, but rather from the sediment removal activities from Sedimentation Ponds 1 and 2 in the winter of 2022.

It is noted that as discussed below, the effectiveness of the flushing cycle to water quality improvement over time was observed in the fourth quarter of 2022 where the boron concentrations at both SS1 and SP2 improved to below the respective trigger concentration.

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#### August 2, 2022:

To address the boron concentrations at SP1 and SP2, a surface water verification monitoring event was completed on August 2, 2022. The verification monitoring for SP1 indicated an acceptable concentration of boron whereas the verification monitoring for SP2 again, indicated a boron concentration greater than its trigger concentration (0.20 mg/L). However, the verification monitoring event for both SP1 and SP2 denoted acceptable biological results indicating that the surface water quality does not pose a detrimental effect to aquatic life and that the surface water was of acceptable quality for continued discharge and that no further verification monitoring was required (refer to **Section 5.3.4**).

In addition, on August 2, 2022, as part of the routine monitoring program, a sample at surface water monitoring station SP3 was collected. The surface water quality at monitoring station SP3 satisfied the relevant trigger concentrations. Samples were again not collected at surface water monitoring stations SS1, SS10, SS16, and SP4 due to there not being flowing conditions to initiate sample collection.

#### August 4, 2022:

A surface water sample was collected at compliance monitoring station SS1, which had water quality that satisfied the relevant trigger concentrations. Samples were again not collected at surface water monitoring stations SS10, SS16, and SP4 due to there not being flowing conditions to initiate sample collection.

#### Fourth Quarter - October 1 to December 31, 2022

#### October 18, 2022:

Surface water samples were collected at compliance monitoring stations SS1, SP2, SP3, and SP4, as well as internal assessment monitoring location SP1. Surface water samples were not collected at background monitoring stations SS10 and SS16 due to there not being flowing conditions to initiate sample collection. The surface water quality at the required monitoring stations generally satisfied the relevant trigger concentrations, with four (4) exceptions. For the surface water sample collected at SS1, the concentrations of nickel (0.092 mg/L), total chromium (0.062 mg/L), and zinc (0.17 mg/L), respectively, were greater than their respective trigger concentrations (0.27 mg/L, 0.024 mg/L, and 0.06 mg/L). Although a notable decrease in concentration than noted in Q2, the concentration of boron (0.39 mg/L) at SP2 was greater than its respective trigger concentration of 0.20 mg/L.

The elevated total nickel, total chromium, and zinc concentrations noted at the location of surface water monitoring station SS1 are interpreted to be as a result of erosional effects due to ongoing rain during sample collection, as well as the amount of precipitation (15.8 mm) that occurred prior to sampling, which fell within 12 hours of sample collection. The visually identifiable turbid waters for the streamflow at SS1 was confirmed through field measurements for turbidity.

The elevated un-ionized ammonia concentration noted at the location of surface water monitoring station SP2 is interpreted to be as a result of the de-watering of Sedimentation Pond 2 that was occurring at the time of sample collection. Sedimentation Pond 2 was being dewatered in preparation for the sediment build-up within the forebays and main portion of the pond to be removed. The de-watering activities would cause some of the

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sediment and suspended algae within the sedimentation pond to become suspended throughout the water body. The visually identifiable turbid waters for the streamflow at SP2 was confirmed through field measurements for turbidity.

Though elevated, the total nickel, total chromium, and zinc concentrations at monitoring station SS1 and the unionized ammonia concentration at SP2, were noted to be within the historical range.

Based on the aforementioned parameter concentrations, verification monitoring was required for SS1 and SP2. The verification monitoring was completed for SS1 and SP2 on November 28, 2022 and January 13, 2023, respectively, and is discussed below.

#### November 28, 2022:

To address the nickel, total chromium, and zinc parameter concentrations at SS1, a surface water verification monitoring event was completed on November 28, 2022. The verification monitoring indicated an acceptable concentration of nickel, total chromium, and zinc at SS1. In addition, the biological results for SS1 indicated that the surface water quality does not pose a detrimental effect to aquatic life and that the surface water was of acceptable quality for continued discharge and that no further verification monitoring was required (refer to **Section 5.3.4**).

#### January 13, 2023:

A verification surface water sample was not collected at monitoring station SP2 during Q4 as there was not a sufficient volume of water within Sedimentation Pond 2 to collect a representative sample following the dewatering and sediment removal activities before the end of Q4. To address the unionized ammonia parameter concentration at SP2, a surface water verification monitoring event was completed in Q1 2023 on January 13, 2023, as there was a sufficient volume of water within the pond to collect a representative sample. The verification monitoring indicated an acceptable concentration of unionized ammonia but a boron concentration greater than its trigger concentration (0.02 mg/L) at SP2. However, the biological results for SP2 indicated that the surface water quality does not pose a detrimental effect to aquatic life and that the surface water was of acceptable quality for continued discharge and that no further verification monitoring was required (refer to Section 5.3.4).

To address the general erosional effects due to significant rainfall, erosion effects related to Street C construction in 2021 that was upstream of SS1 and SP2, and sediment removal activities from Sedimentation Ponds 1 and 2, efforts were completed throughout 2022 and will continue to be undertaken by WM to install additional strawbale check dams at various locations within the surface water drainage network (with focus upstream of SS1 and SP2). It is noted that other sediment control measures have been implemented by WM at the site such as the addition of topsoil and grass seed the interim landfill capped portions of the Expansion Site. For example, throughout October 2022, WM added topsoil and grass seed to the upper interim capped portions of the western sideslope of Cell 2, as well as the western and southern sideslopes of Cell 1. It is expected that as the grass grows, it will reduce the erosional effects that have been observed.

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VOC constituent concentrations were below their relevant laboratory RDLs for the surface water samples collected in 2022.

## 5.4 Landfill Gas Monitoring

Consistent with historical observations, methane gas was not detected within the gas probes in 2022. This observation indicates that landfill gas is not migrating in the subsurface beyond the waste footprint and is therefore, being effectively captured by the landfill gas collection system. Therefore, mitigation measures that would address a potential landfill gas migration in the shallow subsurface beyond the waste footprint are not required. The 2022 landfill gas monitoring results are presented in **Table J-1**, **Appendix J**.

# 6 SITE OPERATIONS SUPPLEMENTAL TESTING & MONITORING

### **6.1 Construction Activities**

### 6.1.1 Existing Site

On October 27, 2022, eight (8) landfill cap repairs were completed on the Existing Site to address Total Hydrocarbon (THC) survey findings. Of the eight (8) repairs, seven (7) were along the eastern sideslope and one (1) was on the western sideslope. These repairs were completed with a geosynthetic bentonite composite layer placed approximately 0.3 m below final cap surface that was then hydrated, and subsequently the clayey soil cap material was replaced and re-compacted. Validation of the ground level THC (as methane) level was completed for the repair locations on November 11, 2022, with one exception which was subsequently repaired again and validated the same day. The ground level THC concentrations were less than 500 ppm indicating that the repairs were successful. Refer to **Volume 4** of the 2022 Annual Report for details regarding the THC Survey monitoring.

### 6.1.2 Expansion Site

The Site was formally approved for expansion on August 5, 2008. During 2022, waste disposal occurred in Cell 4 and Cell 6A of the Expansion Site. Waste disposal in Cell 6A commenced on September 14, 2022, as the landfill liner system was approved to accept waste. Cell 6B pre-excavation activities related to the future construction of the Cell 6B landfill liner system of the Expansion Site commenced in 2022.

Interim cover was placed on the southern, eastern, and western side slopes of Cell 1, the eastern and western side sloped of Cell 2, and the northern and western side slopes of Cell 4 during 2022. In addition, topsoil and seeding was placed on the southern, eastern, and western side slopes of Cell 1 and the eastern and western side sloped of Cell 2.

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### 6.1.3 Groundwater Monitoring Wells

In 2022, groundwater monitoring well nests OW82, OW83, and OW84, along with gas probes GP9 and GP10 were installed at the Site. There were no groundwater monitoring wells or gas probes decommissioned and/or replaced in 2022.

The 2022 monitoring well and gas probe installation/decommissioning status summary is provided in **Appendix M**.

### 6.1.4 On-Site Surface Water Management

Upon completion of cell construction and filling in the Expansion Site, surface water runoff will be managed through Sedimentation Ponds 1, 2, 3, and 4. Surface water runoff originating from areas south of the Existing Site flows toward: 1) Kersey Drain (Brown Creek) to the east; and 2) to the west toward the Van Kessel Drain and enters a municipal drainage tile at a catch basin (SS1), which is situated 60 m east of the western Site boundary. The municipal drainage tile subsequently drains into the discharge ditch for SP2, which ultimately flows to the Gilliand-Geerts Drain 'A', beneath Lambton Road 79.

During the 2022 operating period, maintenance to the on-Site surface water flow system was required, which is detailed in **Appendix R**.

Only water that satisfied the Criteria for On-Site Surface Water Management (D&O, 2008) was used on-Site for dust control and road cleaning purposes in 2022.

### **6.2 MECP Site Inspection Reports**

A MECP Inspector provided inspection summary reports on a semi-annual basis, at a minimum in 2022. The MECP inspection frequency was completed, at a minimum, in accordance with the Waste ECA, as well as in consideration of the MECP's policy regarding field inspections during the COVID Pandemic. The Site inspections were conducted to assess Site operation compliance with the applicable approval documentation. The 2022 MECP Inspection summary reports, received by WM to date, are presented in **Appendix N**. Where action items were required, they were addressed by WM.

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### 6.3 Automobile Shredder Residue

Composite samples of stockpiled ASR material were collected on April 6, and October 5, 2022, for the TCLP parameters listed in Condition 6.52 of the Waste ECA. ASR sampling was not required to be completed during the 2022 Q1 monitoring period. However, a composite sample of stockpiled ASR material that was newly sourced, was collected for laboratory testing on February 2, 2022. Testing results indicated that the ASR was classified as non-hazardous and could be either disposed of in the landfill or utilized as daily cover in accordance with the stipulations of the Waste ECA. Historical analytical data, as well as analytical data related to the 2022 semi-annual (spring and fall) and the February ASR assessments that were completed per Condition 6.52 of the Waste ECA, is presented in **Table K-1**, **Appendix K**. The relevant laboratory Certificates of Analysis, which detail the chemical analytical results for the ASR samples collected in 2022, are included in **Appendix K-2**.

### 6.4 Contaminated Soil

Composite samples of stockpiled contaminated soil were collected during each quarter of 2022 and were submitted for analytical testing of the TCLP parameters listed in Condition 6.57 of the Waste ECA. Testing results indicated that the contaminated soil was classified as non-hazardous and could be either disposed of in the landfill or utilized as daily cover in accordance with the stipulations of the Waste ECA. Historical analytical data, as well as analytical data related to the quarterly contaminated soil assessments that were completed per Condition 6.57 of the Waste ECA, is presented in **Table O-1**, **Appendix O**. The relevant laboratory Certificates of Analysis, which detail the chemical analytical results for the contaminated soil samples collected in 2022, are included in **Appendix O-2**.

## 7 AMBIENT AIR QUALITY MONITORING PLAN

## 7.1 Total Suspended Particulate (TSP) Monitoring

Monitoring of Total Suspended Particulate (TSP) for the 2022 monitoring period was completed as required. Findings of the TSP monitoring program is detailed within **Volume 4** of the 2022 Annual Report.

## 7.2 Volatile Organic Compounds (VOCs) Monitoring

The ambient air volatile organic compound (VOC) monitoring is to be completed during the summer period, which begins on June 21 and ends September 21 of every year in accordance with the Ambient Air Quality Monitoring Plan. Findings of the VOC monitoring program, completed during the 2022 monitoring period, are detailed within **Volume 4** of the 2022 Annual Report.

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### 7.3 Total Hydrocarbon (THC) Landfill Cap Surveys

The landfill final cap surface of the Existing Site is surveyed for Total Hydrocarbon (THC) vapour releases as part of the Ambient Air Quality Monitoring Plan twice annually in the spring and the fall. The landfill final cap surface of the Existing Site was surveyed by RWDI in the spring and fall of 2022 for Total Hydrocarbon (THC) vapour releases, as required. The THC Landfill Cap survey and associated details are presented within **Volume 4** of the 2022 Annual Report.

## 8 NOISE MONITORING PLAN

The Environmental Noise Monitoring Program (Aercoustics, 2007) was implemented in 2009. The survey is required under Condition 13.10 of the Waste ECA. The Noise Monitoring Plan (NMP) report as it relates to the 2022 monitoring period is presented within **Volume 5** of the 2022 Annual Report.

## 9 BEST MANAGEMENT PRACTICES

### **9.1 Dust**

The Best Management Practices Plan (Dust) for the Twin Creeks Environmental Centre was utilized by WM for its operations during 2022. The Dust Inspection and Dispatch Log for each event as filled out by WM are maintained on file.

Complaints related to dust were not received by WM during the 2022 monitoring period.

Details of the Ambient Air Quality Monitoring Program for the Site are presented within **Volume 4** of the 2022 Annual Report.

### 9.2 Litter

WM operated the Site in accordance with its Best Management Practices Plan (Litter) during 2022. The Litter Inspection and Dispatch Logs for each event, as filled out by WM, are maintained on file.

Details related to litter complaints received and the associated response actions(s) by WM during the 2022 monitoring period are outlined in **Section 10 and Appendix P**.

### 9.3 Odour

Odour control is achieved through the Best Management Practices Plan (Odour). The Odour Control Plan was implemented during 2022. Additionally, an odour suppression system is available to manage refuse odours during normal operating procedures as identified by WM.

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Details related to odour complaints received during the 2022 monitoring period are outlined in **Section 10 and Appendix P**.

Further explanation of odour as it relates to the Ambient Air Quality Monitoring Program for the Site is presented within **Volume 4** of the 2022 Annual Report.

### 10 COMPLAINTS

In 2022, WM received a total of 73 complaints (68 for odour, 4 for litter, 1 for trackout of clay soil material, and 1 for the site entrance gates being open during non-operating hours). It is noted that one (1) of the odour complaints and the complaint regarding the open site entrance gates were submitted to WM as one (1) complaint. Of the 73 complaints received, they represented a total of 56 complaint driven events, which occurred on 52 separate days. Of the 56 complaint driven events, 50 of the events were related to odour, four (4) were related to litter, one (1) was related to trackout of clay soil material, and one (1) was related to the site entrance gates being open during non-operating hours. Of the 50 odour events, 18 were documented from 26 discrete physical locations such as a residence or commercial building. The other 32 odour events were transient (driveby) occurrences in which the complainant observed an odour while in transit along a road near to the Site. Further details on these complaint driven events are discussed in **Section Q1.17, Appendix Q, Volume 2.** 

A summary of the complaints and the associated response action for each complaint is presented in **Appendix P**.

## 11 WATER TAKINGS

Reporting of 2022 water takings is required to be completed for Sedimentation Ponds 1 to 4 and the SDL per Ontario Regulation 387/04: Water Taking and Transfer (O. Reg. 387/04), and the Site's PTTW No. 4682-BLJRYJ, dated November 8, 2021. A report indicating water takings during 2022 from the Sedimentation Ponds and the SDL will be submitted to the MECP by March 31, 2023, in accordance with O. Reg. 387/04, and is provided under separate cover, and/or submitted using the online Water Taking Reporting System (WTRS).

In 2022, water was taken from Sedimentation Ponds 1, 2, 3, and 4, as well as pumping station PS6. Water that was taken from the aforementioned ponds was used mainly for dust control for landfill operations. Water was also taken from the aforementioned pumping stations for use as recompacted clayey liner soil conditioning. During 2022, the water taking activities were in compliance with the PTTW limits for the Site, as summarized below.

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	Pond 1	
PTTW Regulatory Components	PTTW Value Limits	2022 Water Taking Values
Max. Taken per Minute (L/min)	4,921	1,064
Max. Hours Taken per Day	24	2.30
Max. Litres Taken Per Day	7,085,520	147,000
Max. Days Taken per Year	105	2
Total Litres Taken in 2022	N/A	279,000
	Pond 2	
PTTW Regulatory Components	PTTW Value Limits	2022 Water Taking Values
Max. Taken per Minute (L/min)	4,921	2,089
Max. Hours Taken per Day	24	3.38
Max. Litres Taken Per Day	7,085,520	423,962
Max. Days Taken per Year	105	39
Total Litres Taken in 2022	N/A	7,570,750
	Pond 3	.,,,,,,,,,,
PTTW Regulatory Components	PTTW Value Limits	2022 Water Taking Values
Max. Taken per Minute (L/min)	4,921	2,089
Max. Hours Taken per Day	24	4.48
Max. Litres Taken Per Day	7,085,520	561,000
Max. Days Taken per Year	105	27
Total Litres Taken in 2022	N/A	6,231,000
	Pond 4	
PTTW Regulatory Components	PTTW Value Limits	2022 Water Taking Values
Max. Taken per Minute (L/min)	4,921	1,064
Max. Hours Taken per Day	24	6.95
Max. Litres Taken Per Day	7,085,520	444,000
Max. Days Taken per Year	105	29
Total Litres Taken in 2022	N/A	5,604,000
	PS6	
PTTW Regulatory Components	PTTW Value Limits	2022 Water Taking Values
Max. Taken per Minute (L/min)	1,325	234
Max. Hours Taken per Day	24	14.10
Max. Litres Taken Per Day	1,907,640	198,000
Max. Days Taken per Year	365	15
Total Litres Taken in 2022	N/A	1,830,000

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## 12 WASTE DISPOSAL INFORMATION

Waste disposal area Cell 6A was constructed in 2022. For reference, design drawings for the Existing and Expansion Sites are presented in **Appendix L**. Details of the 2022 landfill operations are summarized in the relevant sections of **Appendix Q**. **Appendix R** provides the 2022 performance report for the sewage works.

## 13 EAA MONITORING & ANNUAL REPORTING

## 13.1 Mitigation Measures

No mitigation measures related to hydrogeology were required beyond the mitigation built into the Site design.

## 13.2 Monitoring Measures

Monitoring requirements at the landfill evolved in accordance with the EMP as construction progressed in 2022. No changes occurred in 2022 from the previous monitoring period.

## 13.3 Contingency Measures

Based on the groundwater and surface water quality results presented in **Sections 5.2** and **5.3**, the detailed development and implementation of contingency plans is not required.

## 13.4 Annual Reporting

Reporting requirements for the Site are addressed and detailed in **Appendix Q**, per the Waste and Sewage ECAs, as well as the amended PTTW.

## 14 2023 MONITORING PROGRAM

The 2023 Monitoring Program reflects the EMP approved for the Site and incorporates the requirements of the Waste ECA, Sewage ECA, Air ECA, as well as the PTTW. Details of the proposed 2023 Monitoring Program are presented in **Table 13**.

## 15 CONCLUSIONS

Based on the findings presented in this report, the following conclusions are provided.

Overall, generally leachate elevations increased from November 2021 into May 2022 and then decreased
from May 2022 to November 2022. This generalized pattern was expected for 2022 with the short-term
increase of leachate levels in May 2022 resulting from utilizing the leachate volumes within the waste for
irrigation application to the Poplar System beginning in May. From an environmental stewardship
perspective, the leachate generated from within the waste mound of the Existing Site was determined to
be of more suitable use as irrigation liquid during the growing season as opposed to dedicating efforts

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for off-Site disposal and treatment between late winter and spring. This storage practice began in 2019 and will continue into the future such that there is sufficient weak-strength leachate volume for irrigation purposes during the growing season. There was a short-term slight increase in leachate elevation noted at a number of the leachate maintenance holes from May 2022 to November 2022, however, the leachate elevations were noted to be within their historical ranges.

- Environmental compliance at the Site as it relates to the Existing Site relies on groundwater and surface water quality monitoring, which once again verified at the Site in 2022. As an additional guidance tool toward managing leachate volumes in the waste cells a hydraulic gradient assessment of leachate compared to nearby shallow groundwater is utilized to further protect local water resources. May 2022 leachate levels were reflective of leachate storage for utilization onto the Poplar. However, by November 2022, the leachate elevation data within the waste of the Existing Site indicated that groundwater flow was inward toward the waste at Cells 3, 5, 7, 8, 9, 10, 11, 12, as well as the southern and eastern portions of the South Cell.
- In 2022, leachate elevations within the eastern portion of Cell 3S (at MH3SA and MH3SB), Cell 6, the western portion of the South Cell (at OW22A-10 and OW53-10), and the West Cell (Sump) were higher in elevation compared to groundwater during the May monitoring event. Leachate elevations within the western portion of Cell3S (at MH3SC), the northern portion of Cell 4 (at MH4B), Cell 6, the western portion of the South Cell (at OW22A-10 and OW53-10), and the West Cell (Sump) were higher in elevation compared to groundwater during the November monitoring event. However, acceptable water quality was demonstrated at the Site during 2022 and therefore, the elevated leachate elevations that showed the potential for outward migration were not negatively affecting the water resources at the Site. Also, leachate seeps were not observed along the waste side slopes of the Existing Site in 2022. Continued leachate management of the leachate via automated pumping is expected to continue to reduce the leachate mound in these cells overtime.
- During May and November 2022, leachate elevation in the West Cell (Sump) was greater than the surrounding natural ground surface elevation. However, the leachate elevation in the West Cell (Sump) was lower than its historical high in May 2019 by 2.71 m in May 2022 and by 2.58 m in November 2022, through leachate transfer to the Expansion Site Leachate Equalization Tank. The noted leachate elevation within the Sump does not represent a concern based on acceptable surface water quality at compliance stations and no visible leachate seeps on the sideslope. Continued leachate extraction from the West Cell over time is expected to further reduce the potential for outward leachate seepage in this area. An evaluation will be completed in 2023 to assess if the leachate level can be further lowered in the Sump without compromising the leachate extraction pump.
- Overall, between May 2021 and May 2022 there was a slight increase in the calculated theoretical total leachate volume stored above the local groundwater table (31,904 m³) and within the waste (13,284 m³) of the Existing Site. Between November 2021 and November 2022 there was a decrease in the calculated theoretical total leachate volume stored above the local groundwater table (34,275 m³) and within the waste (36,911 m³) of the Existing Site. It is noted that these volumes are only theoretical in nature and are used as guidance information for year over year overall performance evaluation. For example, the total leachate storage in the Existing Site decreased from May to November 2022 (a period of time approximately 2 weeks longer than the irrigation season) by 28,572 m³, however, during that time approximately 7,173 m³ was measured to have been extracted from the Existing Site. During this

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- timeframe, the pumping stations in many of the weaker strength leachate waste cells were routinely drawn down to elevations where the leachate elevation was too low to safely operate the pumps.
- The trigger mechanism for implementation of groundwater contingency measures for the Expansion Site is the loss of hydraulic containment of the landfill waste footprint. During 2022, the leachate target level for each pumping station of the Expansion Site as noted in Condition 14.1 of the Waste ECA was satisfied.
- The operation of the primary leachate collection system is designed to maintain a liquid level of no more than 300 mm (or 0.30 m) above the primary clay liner (bottom of the landfill. One (1) exceedance was noted, on September 29, 2022, where the leachate elevation exceeded the 0.3 m of head by 0.01 m. It is noted that periodic elevated leachate occurrences are described in the D&O Report after precipitation or snow melt events generate more leachate than the pumps can extract and a temporary increase of the leachate head under such circumstances is not considered a non-compliance issue with the Waste ECA.
- Between January 1 and December 31, 2022, a total of 54,626.76 m³ of leachate was managed, of which 47,454.19 m³ of leachate was removed and transported off-Site for treatment and disposal at the Chatham Water Pollution Control Plant, while 7,172.57 m³ was irrigated onto the Poplar System during the 2022 growing season.
- Of the 54,626.8 m³ noted above, the approximate breakdown of leachate source location between the Existing Site and the Expansion Site is 24% (13,132.37 m³) and 76% (41,494.39 m³), respectively. This breakdown is based on the leachate source distribution which incorporates the separate approximate volumes of leachate extracted from the Existing Site (38% of area) and Expansion Site (62% of area) for either off-Site disposal of stronger strength leachate (i.e. Expansion Site leachate) or storage of weaker strength leachate (i.e. Existing Site leachate) for use as irrigation liquid to the Poplar System during the 2022 growing season.
- Considering a north-south groundwater drainage divide at the Site, shallow groundwater movement within the active aquitard was in an easterly and westerly direction away from the divide with local variations as a result of surficial feature influences, such as ditches, waste cells, excavations, ponds, berms, and the leachate management systems. Groundwater movement within the interstadial silt and sand also typically moves toward the east and west, away from the drainage divide. Influences from the excavation of the Expansion Site as well as landfilling of waste in the Expansion Site were noted to the south (at monitoring well OW79) where the groundwater in the active aquitard, as well as the interstadial silt and sand was induced toward the north. The groundwater pressures within the interface aquifer have also appeared to be affected by the removal of the overlying soil mass, whereby pressures were observed to be slightly decreasing in recent years, and are expected to return to pre-existing conditions as the soil mass is replaced with waste mass. As expected, monitoring wells OW39A-26, OW79-26, OW80-27, and OW81-27 are showing a fluctuating and increasing trend since about 2018 as a significant waste mass has been progressively landfilled into the southern half of the Expansion Site. Groundwater movement within the interface aquifer is in a southwesterly direction.

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- The groundwater monitoring results for the active aquitard, the interstadial silt and sand, as well as the interface aquifer satisfied the relevant PLIL and SLIL trigger concentrations. One (1) exception occurred during the spring monitoring event for the PLIL parameter chloride at OW81-7, which as discussed in **Section 5.2.3.2** is not landfill related. It is noted that based on road salting influences in this area, approval to remove chloride as a PLIL parameter at this location was issued by the MECP on August 18, 2022.
- At monitoring wells OW69-5 (Active Aquitard), and OW46-7 (Interstadial Silt and Sand), the concentrations of boron show infrequent spikes, but concentrations are less than the respective trigger concentration at each location. At this time the boron concentrations are not a concern as it is expected the more mobile parameter chloride would be observed first if the source of the boron were leachate. It is likely that, similar to observations for monitoring well OW58-14, the bentonite seal for these locations is likely moving into the filter screen material of the monitoring well, and as such may require decommissioning and replacement in the future depending on chemical results. Overall, groundwater quality did not show an unacceptable landfill leachate or operations effect in 2022.
- The routine quarterly surface water monitoring results satisfied the relevant trigger concentrations, with seven (7) exceptions. The exceptions are discussed in detail **in Section 5.3.5**, with verification results indicating acceptable chemical and biological results and no further verification monitoring was required. Overall, surface water quality did not show an unacceptable landfill leachate or operations effect in 2022.
- The annual spring biomonitoring showed that there was not a potential for detrimental effects to aquatic life in the discharge water and was acceptable for continued discharge.
- Water takings at the Site from the Sedimentation Ponds were used as a dust suppressant related to landfill operations. The water taking activities in 2022 satisfied the requirements of the Site's PTTW.
   Documentation will be submitted to the MECP, as required by the PTTW, under separate cover.
- No methane gas was detected within the gas probes in 2022. Therefore, mitigation measures that would
  address a potential landfill gas migration in the shallow subsurface beyond the waste footprint are not
  required.
- In 2022, WM received a total of 73 complaints (68 for odour, 4 for litter, 1 for trackout of clay soil material, and 1 for the site entrance gates being open during non-operating hours). It is noted that one (1) of the odour complaints and the complaint regarding the open site entrance gates were submitted to WM as one (1) complaint. Of the 73 complaints received, they represented a total of 56 complaint driven events, which occurred on 52 separate days. Of the 56 complaint driven events, 50 of the events were related to odour, four (4) were related to litter, one (1) was related to trackout of clay soil material, and one (1) was related to the site entrance gates being open during non-operating hours. Of the 50 odour events, 18 were documented from 26 discrete physical locations such as a residence or commercial building. The other 32 odour events were transient (drive-by) occurrences in which the complainant observed an odour while in transit along a road near to the Site.

WASTE MANAGEMENT OF CAN RWDI#2202861.01 March 1, 2023



## 16 CLOSURE

We trust that this 2022 Fourth Quarter and Annual Monitoring Report for the Twin Creeks Environmental Centre is satisfactory. Should there be any questions or comments, please contact us.

Yours very truly,

**RWDI AIR Inc.** 

Jeff Cleland, P.Eng Environmental Engineer

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Attach.

Brent J. Langille, B.Sc., P.Geo. Technical Director | Principal

RWDI#2202861.01 March 1, 2023



# 17 STATEMENT OF QUALIFICATIONS AND LIMITATIONS

This Report has been prepared for a specific purpose and use, as outlined within the Report. The scope of the undertaking was initially provided in a proposal submitted by RWDI AIR, Inc. (RWDI) to Waste Management of Canada Corporation. The proposal (subject to any documented scope changes requested by Waste Management of Canada Corporation) constitutes an agreement between RWDI and Waste Management of Canada Corporation.

RWDI relied in part, upon the data, information, specifications and documentation (Data) provided by Waste Management of Canada Corporation as well as third parties. It is assumed by RWDI that the Data provided are complete and accurate. RWDI was not retained to, nor has it conducted any independent verification of the accuracy, completeness or suitability of the Data. As such, RWDI assumes no liability for losses, damages, or claims of any nature arising from inaccurate, incomplete or unsuitable Data provided on this project. Waste Management of Canada Corporation by receipt of this Report agrees to indemnify and hold harmless RWDI with respect thereto.

It is important that the reader of this Report, recognize that subsurface, environmental and/or geotechnical conditions may vary geographically and temporally. This is a natural phenomenon, which is not fully accommodated in the limited testing conducted by RWDI. In addition, the analysis of the collected data, by necessity, incorporates simplifying assumptions of site conditions and analytical solutions that assume uniformity in site conditions. The opinions, conclusions, and recommendations contained within the Report therefore represent RWDI's professional judgment in-light of these limitations.

This Report is to be considered confidential and is for the sole use of Waste Management of Canada Corporation. As such, the Report shall not be relied upon by third parties, except where agreed in writing between RWDI and Waste Management of Canada Corporation; where required by law; or where used for governmental review. RWDI accepts no responsibility, and denies any liability whatsoever, to parties other than Waste Management of Canada Corporation who may obtain access to the Report, for any injury, loss, or damage suffered by such parties arising from their use of, reliance upon, decisions or actions based on the Report or any of its contents, except to the extent where those parties have obtained prior written consent of RWDI to use and rely upon the Report and its contents. Any damages arising from improper use of the Report or parts thereof shall be borne by the party making such use.

This statement of Qualifications and Limitations is attached to and forms part of the Report and any use of the Report are subject to the terms thereof.



## **TABLES**

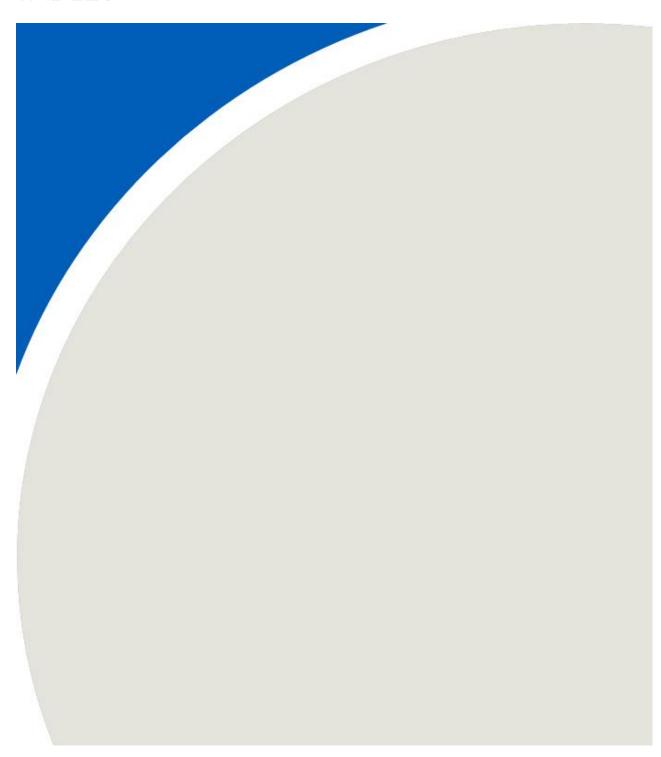


Table 1 Monitoring Schedule - 2022 Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Task	Monitoring Locations	Monitoring Dates	Notes
	First Quarter Monitoring Per	iod (January 1 to March 31, 2022)	
Precipitation Event	SS1, SS10, SS16, SS19 (new), SP1, SP2, SP3, SP4	February 17, 2022 - Routine monitoring for February 16, 2022, precipitation event.	SS19 not monitored since the compost facility is not yet constructed. SS10 not monitored due to no flow conditions.
Monitoring/Sampling res		May 4, 2022 - Verification monitoring event based on the results for the February 17, 2022 routine monitoring event.	
Biomonitoring	SS1	May 4, 2022 - Verification monitoring event based on the results for the February 17, 2022 routine monitoring event.	
Leachate Liquid Level Measurements	PS1, PS3, PS5, PS7(new), MH3S, MH4, MH5, MH6, MH7, MH8, MH9, MH10, MH11, MH12, MH16, MH17, MH18, OW22A-10, OW51A-15, OW53-10, Sump	Semi-annually in May and November. Daily during operation for PS1, PS3, and PS5.	PS7 was not monitored as it was not yet constructed.
Leachate Sampling	Equalization Tank	Quarterly, and Semi-Annually in May and November. A quarterly sample was collected on January 19, 2022.	
Secondary Drainage Layer (SDL) Liquid Levels	PS2, PS4, PS6, PS8(new)	Monthly - January 5, February 2, and March 1, 2022 - PS2, PS4, and PS6	PS8 was not monitored as it was not yet constructed.
Gas Monitoring	GP1A, GP2, GP3, GP4, GP5, GP6, GP7, GP8, GP9 (new), GP10 (new)	Monthly gas monitoring from November to April and in July. Perimeter gas probes were monitored on January 5, February 3, and March 1, 2022.	Gas probes GP9, and GP10 were not yet installed.
	Poplar System I	Monitoring Program	
Precipitation Event Surface Water	SS14A, SS14B, and SS15A	February 17, 2022 - Routine monitoring for February 16, 2022, precipitation event.	
Monitoring/Sampling			
		y Monitoring Program	
Total Suspended Particulate - Dust	Stations West, Northeast, and Southeast	Every sixth day - NAPS Schedule (October 1 to May 31) Every third day - NAPS Schedule (June 1 to September 30)	
	Noise Mon	itoring System	
Noise monitoring	Stations - M1, M2, M3, and M4	Ongoing - Quarterly Reporting	
	· · · · · · · · · · · · · · · · · · ·	al Monitoring	
Contaminated Soil	Landfill Daily Cover/Disposed Material	Quarterly, if utilized: January 11, 2022.	
Automobile Shredder Residue (ASR)	Landfill Daily Cover/Disposed Material	Semi-Annually (Spring and Fall) if utilized: February 2, 2022.	

- (new) denotes monitoring station/wells to be installed per the Environmental Monitoring Plan (EMP) dated December 20, 2007.
   OW71A-5* denotes groundwater well monitored at the request of the Public Liaison Committee.

Task	Monitoring Locations	Monitoring Dates	Notes
	Second Quarter Monitoring P	Period (April 1 to June 30, 2022)	
Precipitation Event	SS1, SS10, SS16, SS19 (new), SP1, SP2, SP3, SP4	May 4, 2022 - Routine monitoring for May 3, 2022 precipitation event.	SS19 not monitored since the compost facility is not yet constructed.
Surface Water Monitoring/Sampling	SS1, SP2	May 16, 2022 - Verification monitoring event based on the results for the May 4, 2022 routine monitoring event.	
	SS1, SS10, SS16, SS19 (new), SP1, SP2, SP3, SP4	May 4, 2022 - Routine monitoring for May 3, 2022 precipitation event.	SS19 not monitored since the compost facility is not yet constructed.
Biomonitoring	SS1, SP2	May 16, 2022 - Verification monitoring event based on the results for the May 4, 2022 routine monitoring event.	
Leachate Liquid Level Measurements	PS1, PS3, PS5, PS7(new), MH3S, MH4, MH5, MH6, MH7, MH8, MH9, MH10, MH11, MH12, MH16, MH17, MH18, OW22A-10, OW51A-15, OW53-10, Sump	Semi-annually in May and November: May 2, 2022. Daily during operation for PS1, PS3, and PS5.	PS7 was not monitored as it was not yet constructed.
	PS1, PS3, PS5, PS7(new), South Fill Area (MH18), West Central Fill Area (Sump), Central Fill Area (Composite of MH3S, MH4, MH5, MH6, MH7, MH8, MH9, MH10, MH11, MH12)	Annually in May: May 6 and 10, 2022.	PS7 was not monitored as it was not yet constructed.
	Equalization Tank	Semi-Annually sampled in May and November: May 24, 2022.	
Secondary Drainage Layer (SDL) Liquid Levels	PS2, PS4, PS6, PS8(new)	Monthly - April 8, May 2, and June 8, 2022 - PS2, PS4, and PS6.	PS8 was not monitored as it was not yet constructed.
Groundwater Liquid Level Measurements	ACTIVE AQUITARD  OW16-6, OW17-4, OW40B-4r, OW54A-4, OW56-4, OW57-4, OW58-6, OW59-6, OW60-4, OW67-4, OW68-5, OW69-5, OW70B-5, OW71A-5*, OW72-6, OW73-6, OW79-5, OW80-3, OW81-5, OW82(new), OW83(new), OW84(new)  INTERSTADIAL SILT AND SAND  OW16-7, OW40A-7, OW46-7, OW47-6, OW54-10, OW57-15, OW58-17, OW60-8, OW67-11, OW72-10, OW73-9, OW79-7, OW80-6, OW81-7, OW82(new), OW83(new), OW84(new)  INTERFACE AQUIFER  OW17-30, OW19-29, OW39-26, OW40A-28, OW49-29, OW60-25, OW79-26, OW80-27, OW81-27, OW82(new), OW83(new), OW84(new)	Semi-annually in May and November: May 2, 2022.	Well nests OW82, OW83, and OW84 were not monitored as they were not yet constructed.
Piezometer Liquid	PS1, PS2, PS3	Semi-annually in May and November: May 2 , 2022.	
Level Measurements			
ACTIVE AQUITARD OW16-6, OW17-4, OW40B-4r, OW54A-4, OW56-4, OW57-4, OW58-6, OW59-6, OW60-4, OW67-4, OW68-5, OW69-5, OW70B-5, OW71A-5*, OW72-6, OW73-6, OW79-5, OW80-3, OW81-5, OW82(new), OW83(new), OW84(new) INTERSTADIAL SILT AND SAND OW16-7, OW40A-7, OW46-7, OW47-6, OW54-10, OW57-15, OW58-14, OW60-8, OW67-11, OW72-10, OW73-9, OW79-7, OW80-6, OW81-7, OW82(new), OW84(new) INTERFACE AQUIFER OW19-29, OW39-26, OW49-29, OW79-26, OW80-27, OW81-27, OW82(new), OW83(new), OW83(new), OW84(new), Cemetery Well		Semi-annually in May and November: May 3, 4, 5 and 6, 2022.	Well nests OW82, OW83, and OW84 were not monitored as they were not yet constructed.
		June 2, 2022 - OW81-7 verification monitoring.	
Gas Monitoring	GP1A, GP2, GP3, GP4, GP5, GP6, GP7, GP8, GP9 (new), GP10 (new)	Monthly gas monitoring from November to April and in July. Perimeter gas probes were monitored on April 27, 2022.	Gas probes GP9, and GP10 are not yet installed.
		onitoring Program	
Precipitation Event Surface Water Monitoring/Sampling	SS14A, SS14B, and SS15A	May 4, 2022 - Routine monitoring for May 3, 2022 precipitation event.	
		Monitoring Program	
Total Suspended Particulate - Dust	Stations West, Northeast, and Southeast	Every sixth day - NAPS Schedule (October 1 to May 31) Every third day - NAPS Schedule (June 1 to September 30)	
Volatile Organic Compounds	Upwind/downwind Monitoring	June 21 to September (5 sets): June 24, 2022	
Total Hydrocarbon Landfill Cap Survey	Final Capped Areas	Between the Spring and Fall: June 15, 2022	
		oring Program	
Noise Monitoring	Stations - M1, M2, M3, and M4  Operationa	Ongoing - Quarterly Reporting.  I Monitoring	
Contaminated Soil	Landfill Daily Cover/Disposed Material	Quarterly, if utilized: April, 6 2022.	
	Landfill Daily Cover/Disposed Material	Semi-Annually (Spring and Fall), if utilized: April 6, 2022.	
Automobile Shredder		and the second of the second o	

- (new) denotes monitoring station/wells to be installed per the Environmental Monitoring Plan (EMP) dated December 20, 2007.
   OW71A-5* denotes groundwater well monitored at the request of the Public Liaison Committee.

Task	Monitoring Locations	Monitoring Dates	Notes
	Third Quarter Monitoring Peri	od (July 1 to September 30, 2022)	
		onitoring System	SS19 not monitored since the compost facility is
	SS1, SS10, SS16, SS19(new), SP1, SP2, SP3, SP4	July 20, 2022 - Routine monitoring for July 19, 2022 precipitation event.	not yet constructed. SS1, SS10, SS16, SP3, and SP4 not monitored due to no flow conditions.
Precipitation Event Surface Water	SP3	August 2, 2022 - Routine monitoring for August 1, 2022 precipitation event.	SS1, SS10, SS16, and SP4 not monitored due to no flow conditions.
Monitoring/Sampling	SS1	August 4, 2022 - Routine monitoring for August 3, 2022 precipitation event.	SS10, SS16, and SP4 not monitored due to no flow conditions.
	SP1 and SP2	August 2, 2022 - Verification monitoring event based on the results for the July 20, 2022 routine monitoring event.	
Biomonitoring	SP1 and SP2	August 2, 2022 - Verification monitoring event based on the results for the July 20, 2022 routine monitoring event.	
Leachate Level Measurements	PS1, PS3, PS5, PS7, MH3S, MH4, MH5, MH6, MH7, MH8, MH9, MH10, MH11, MH12, MH16, MH17, MH18, OW22A-10, OW51A-15, OW53-10, Sump	Semi-annually in May and November. Daily during operation for PS1, PS3, PS5, and PS7.	PS7 was monitored beginning in September.
Leachate Sampling	Equalization Tank	Quarterly, and semi-annually in May and November. A quarterly sample was collected on July 25, 2022.	
Secondary Drainage Layer (SDL) Liquid Levels	PS2, PS4, PS6, PS8	Monthly - July 8, August 8, and September 1, 2022	PS8 was monitored beginning in September.
Gas Monitoring	GP1A, GP2, GP3, GP4, GP5, GP6, GP7, GP8, GP9, GP10	Monthly gas monitoring from November to April and in July. Perimeter gas probes were monitored on July 28, 2022.	
	Poplar System M	lonitoring Program	
Soil Conductivity	Poplar System	Weekly during the irrigation season. Began on May 11, 2022.	
Soil Sampling	Poplar System	Annually in September. Completed on September 22, 2022.	
Visual Assessments	Poplar System	Annually in September. Completed on September 22, 2022.	
Undergrowth and Ponding Inspections, Odour Monitoring	Poplar System	Monthly during irrigation.	
Leachate Liquid Level Measurements	MH16, MH17, MH18, OW22A-10, OW53-10	Semi-annually in May and November.	Monitoring not required as part of the Q3 monitoring period.
Leachate Sampling	PS Holding Tank  Monthly during the irrigation season: July 7, August 9, and September 1, 2022.		
Leaf Tissue Analysis	Poplar System	Annually in September. Completed on September 22, 2022.	
Stem Core & Root Samples, Root Depth	Poplar System	Annually in September. Completed on September 22, 2022.	
Precipitation Event	SS14A, SS14B, and SS15A	August 4, 2022 - Routine monitoring for August 3, 2022 precipitation event.	SS14A and SS14B not monitored due to no flow conditions.
Surface Water Monitoring/Sampling	SS14B	August 30, 2022 - Routine monitoring for August 29, 2022 precipitation event.	SS14A not monitored due to no flow conditions.
	SS14A	September 28, 2022 - Routine monitoring for September 27, 2022 precipitation event.	SS14A and SS15A not monitored due to no flow
Storm Event Surface	SS14A, SS14B, and SS15A	Two (2) events during the irrigation season and after a storm event (>25 mm in 24 hrs): August 4, 2022 - Storm event monitoring for August 3, 2022 precipitation event.	conditions.
Water Monitoring	SS14A, SS14B, and SS15A	September 28, 2022 - Storm event monitoring for September 27, 2022 precipitation event.	
		/ Monitoring Program	'
Total Suspended Particulate - Dust	Stations West, Northeast, and Southeast	Every sixth day - NAPS Schedule (October 1 to May 31) Every third day - NAPS Schedule (Upp 1 to September 30)	
Volatile Organic Compounds	Upwind/downwind Monitoring	(June 1 to September 30)  June 21 to September (5 sets): July 26, August 12, August 25, and September 9, 2022.	
Total Hydrocarbon Landfill Cap Survey	Final Capped Areas	Between the Spring and Fall: October 3, 2022.	
	Noise Moni	toring System	
Noise Monitoring	Stations - M1, M2, M3, and M4	Ongoing - Quarterly Reporting  al Monitoring	
Contaminated Soil	Landfill Daily Cover/Disposed Material	Quarterly, if utilized: July 6, 2022.	
Automobile Shredder Residue	Landfill Daily Cover/Disposed Material	Semi-Annually (Spring and Fall), if utilized:	Monitoring not completed during the 2022 third quarter monitoring period.

- 1) (new) denotes monitoring station/wells to be installed per the Environmental Monitoring Plan (EMP) dated December 20, 2007.
- 2) OW71A-5* denotes groundwater well monitored at the request of the Public Liaison Committee.

	SS1, SS10, SS16, SS19(new), SP1, SP2, SP3, SP4	Monitoring Dates  I (October 1 to December 31, 2022)  October 18, 2022 - Routine monitoring for October 17,					
Precipitation Event	SS1, SS10, SS16, SS19(new), SP1, SP2, SP3, SP4						
Precipitation Event		October 18. 2022 - Routine monitoring for October 17.	SS19 not monitored since the compost facility is				
		2022 precipitation event.	not yet constructed. SS10 and SS16 not monitored due to no flow conditions.				
Monitoring/Sampling		November 28, 2022 - Verification monitoring event based on the results for the October 18, 2022 routine monitoring event.	SP2 not monitored due to Sedimentation Pond 2 not having sufficient volume to initiate sample collection.				
S		January 13, 2023 - Verification monitoring event based on the results for the October 18, 2022 routine monitoring event.					
Biomonitoring		November 28, 2022 - Verification monitoring event based on the results for the October 18, 2022 routine monitoring event.					
Significant		January 13, 2023 - Verification monitoring event based on the results for the October 18, 2022 routine monitoring event.					
Ecacinate Eigala Ecvel		Semi-annually in May and November: November 1, 2022. Daily during operation for PS1, PS3, and PS5.					
Leachate Sampling	·	Quarterly, and semi-annually in May and November. A quarterly sample was collected on November 7, 2022.					
Secondary Drainage Layer (SDL) Liquid Levels	PS2, PS4, PS6, PS8	Monthly - October 13, November 1, and December 7, 2022					
Groundwater Liquid Level Measurements	OW16-6, OW17-4, OW40D-4, OW54A-4, OW56-4, OW57-4, OW58-6, OW59-6, OW60-4, OW67-4, OW68-5, OW69-5, OW70B-5, OW71A-5*, OW72-6, OW73-6, OW79-5, OW80-3, OW81-5, OW82-5, OW83-5, OW84-6 INTERSTADIAL SILT AND SAND OW16-7, OW40A-7, OW46-7, OW47-6, OW54-10, OW57-15, OW58-14, OW58-17,	Semi-annually in May and November: November 2, 2022.					
9   	OW60-8, OW67-11, OW72-10, OW73-9, OW79-7, OW80-6, OW81-7, OW82-14, OW83-9, OW84-11 INTERFACE AQUIFER OW17-30, OW19-29, OW39A-26, OW40A-28, OW49-29, OW60-25, OW79-26, OW80-27, OW81-27, OW82-28, OW83-29, OW84-31						
Piezometer Liquid Level Measurements	PS1, PS2, PS3	Semi-annually in May and November: November 1, 2022.					
0		Semi-annually in May and November: November 1 and 2, 2022.					
1   I	OW16-7, OW40A-7, OW46-7, OW47-6, OW54-10, OW57-15, OW58-14, OW60-8, OW67- 11, OW72-10, OW73-9, OW79-7, OW80-6, OW81-7, OW82-14, OW83-9, OW84-11 INTERFACE AQUIFER OW19-29, OW39-26, OW49-29, OW79-26, OW80-27, OW81-27, OW82-28, OW83-29, OW84-31, Cemetery Well						
Gas Monitoring		Perimeter gas probes monitored on November 17 and December 1, 2022					
		onitoring Program	00444				
Precipitation Event Surface Water		October 18, 2022 - Routine monitoring for October 17, 2022 precipitation event.	SS14A not monitored due to no flow conditions.				
Monitoring/Sampling		October 19, 2022 - Routine monitoring for October 18, 2022 precipitation event.					
	Ambient Air Quality Monitoring Program						
Total Suspended Particulate - Dust		Every sixth day - NAPS Schedule (October 1 to May 31) Every third day - NAPS Schedule					
2 300		(June 1 to September 30)					
	Noise Monitoring Program						
Noise Monitoring		Ongoing - Quarterly Reporting					
	Omeustions	l Monitoring					
Containinated 3011	Landfill Daily Cover/Disposed Material	Quarterly (if utilized), October 5, 2022					
Automobile Shredder Residue	Landfill Daily Cover/Disposed Material	Semi-Annually (Spring and Fall), if utilized: October 5, 2022					

- 1) (new) denotes monitoring station/wells to be installed per the Environmental Monitoring Plan (EMP) dated December 20, 2007.
  2) OW71A-5* denotes groundwater well monitored at the request of the Public Liaison Committee.

Table 2 Leachate Elevation Comparison Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Monitoring Location	Groundwater Monitoring Wells Compared	Existing Ground Surface Elevation (m ASL)	Exisiting Bottom of Ditch Elevation (m ASL)	Current Local Groundwater Elevation (m ASL)	Leachate Elevation (m ASL)	Leachate Elevation Compared to Ground Surface Elevation	Leachate Elevation Compared to Groundwater Elevation
			May 2, 2022				
			Cell 3S				
MH3SA	OW17-4 and OW56-4	240.71	240.68	239.58	240.09	LOWER	HIGHER
MH3SB	OW17-4 and OW56-4	240.46	240.31	239.58	239.99	LOWER	HIGHER
MH3SC	OW7-5	239.66	239.42	236.85	235.57	LOWER	LOWER
MH3SD	OW7-5	239.87	239.93	236.85	235.37	LOWER	LOWER
			Central Fill Area				
Sump	OW7-5 and OW8-5	240.81	240.28	237.14	242.43	HIGHER	HIGHER
OW51A-15	OW17-4	240.24	239.68	239.72	236.30	LOWER	LOWER
MH4A	OW17-4 and OW69-5	240.33	239.71	239.48	238.50	LOWER	LOWER
MH4B	OW57-4	240.95	240.17	239.60	239.46	LOWER	LOWER
MH5A	OW58-6	241.51	240.78	240.88	239.73	LOWER	LOWER
MH6A	OW65-4/OW65A-4	241.90	241.20	237.34	240.16	LOWER	HIGHER
MH7A	OW73-6	242.07	241.34	240.68	237.15	LOWER	LOWER
MH8B	OW74-6	242.54	242.46	239.33	239.22	LOWER	LOWER
MH9A	OW72-6	242.33	241.89	240.90	239.24	LOWER	LOWER
MH10	OW74-6	241.80	241.43	239.33	236.68	LOWER	LOWER
MH11A	OW54A-4	242.34	241.94	241.90	238.73	LOWER	LOWER
MH12A	OW66-4	241.90	241.37	241.79	236.69	LOWER	LOWER
MH12B	OW66-4	241.90	241.37	241.79	236.55	LOWER	LOWER
			South Cell				
MH16	OW63A-6	239.53	238.49	238.12	238.03	LOWER	LOWER
MH17	OW63A-6	239.12	238.07	238.12	238.09	LOWER	LOWER
MH18	OW63A-6	238.84	238.18	238.12	237.92	LOWER	LOWER
OW22A-10	OW6-4	239.38	238.76	238.11	238.43	LOWER	HIGHER
OW53-10	OW44-5 and OW64-4/OW64A-4	239.47	238.45	236.77	238.53	LOWER	HIGHER
			Expansion Site Cel				
PS1	OW38-6	240.88		236.73	227.00	LOWER	LOWER
			Expansion Site Cel				
PS3	OW38-6	240.18		236.73	226.65	LOWER	LOWER
205	OWDO 6	0.40.70	Expansion Site Cel		227.05	1.014/50	LOWED
PS5	OW38-6	240.73		236.73	227.05	LOWER	LOWER

**Notes:** 1) Leachate elevations from May 2, 2022.

²⁾ m ASL denotes meters above sea level.

³⁾ *Italics* denotes historic information used to calculate data.

⁴⁾ **Bold and italics** denotes elevation presented is the maintenance hole bottom elevation as it was dry at the time of monitoring.

⁵⁾ For waste cells with two maintenance holes, the maintenance hole farthest into the waste cell was utilized for comparison.

⁶⁾ Revised measuring point elevations for Existing Site manholes and leachate wells, as well as updated ground surface elevations reflect the findings of an updated survey (2016).

⁷⁾ Existing Ground Surface Elevation and Existing Bottom of Ditch Elevation for MH16, MH17, and MH18 remeasured on November 3, 2022.

Table 2 Leachate Elevation Comparison Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Monitoring Location	Groundwater Monitoring Wells Compared	Existing Ground Surface Elevation (m ASL)	Exisiting Bottom of Ditch Elevation (m ASL)	Current Local Groundwater Elevation (m ASL)	Leachate Elevation (m ASL)	Leachate Elevation Compared to Ground Surface Elevation	Leachate Elevation Compared to Groundwater Elevation
			November 1, 202	22			,
			Cell 3S				
MH3SA	OW17-4 and OW56-4	240.71	240.68	238.81	238.77	LOWER	LOWER
MH3SB	OW17-4 and OW56-4	240.46	240.31	238.81	238.52	LOWER	LOWER
MH3SC	OW7-5	239.66	239.42	236.85	237.15	LOWER	HIGHER
MH3SD	OW7-5	239.87	239.93	236.85	236.84	LOWER	LOWER
			Central Fill Area				
Sump	OW7-5 and OW8-5	240.81	240.28	237.14	242.56	HIGHER	HIGHER
OW51A-15	OW17-4	240.24	239.68	239.24	236.58	LOWER	LOWER
MH4A	OW17-4 and OW69-5	240.33	239.71	239.00	238.38	LOWER	LOWER
MH4B	OW57-4	240.95	240.17	238.80	239.37	LOWER	HIGHER
MH5A	OW58-6	241.51	240.78	239.18	236.79	LOWER	LOWER
MH6A	OW65-4/OW65A-4	241.90	241.20	237.34	240.23	LOWER	HIGHER
MH7A	OW73-6	242.07	241.34	240.08	236.60	LOWER	LOWER
MH8B	OW74-6	242.54	242.46	239.33	238.90	LOWER	LOWER
MH9A	OW72-6	242.33	241.89	240.89	237.75	LOWER	LOWER
MH10	OW74-6	241.80	241.43	239.33	236.81	LOWER	LOWER
MH11A	OW54A-4	242.34	241.94	240.16	239.81	LOWER	LOWER
MH12A	OW66-4	241.90	241.37	241.79	236.83	LOWER	LOWER
MH12B	OW66-4	241.90	241.37	241.79	236.86	LOWER	LOWER
			South Cell				
MH16	OW63A-6	239.53	238.49	238.12	238.07	LOWER	LOWER
MH17	OW63A-6	239.12	238.07	238.12	238.07	LOWER	LOWER
MH18	OW63A-6	238.84	238.18	238.12	238.08	LOWER	LOWER
OW22A-10	OW6-4	239.38	238.76	238.24	238.47	LOWER	HIGHER
OW53-10	OW44-5 and OW64-4/OW64A-4	239.47	238.45	235.30	238.36	LOWER	HIGHER
			Expansion Site Cel	l1			
PS1	OW38-6	240.88		236.73	227.08	LOWER	LOWER
	·		Expansion Site Cel		•	•	
PS3	OW38-6	240.18		236.73	227.23	LOWER	LOWER
			Expansion Site Cel	14			
PS5	OW38-6	240.73		236.73	227.20	LOWER	LOWER
			Expansion Site Cel	16			
PS7	OW38-6	239.41		236.73	228.16	LOWER	LOWER

**Notes:** 1) Leachate elevations from November 1, 2022.

- 2) m ASL denotes meters above sea level.
- 3) *Italics* denotes historic information used to calculate data.
- 4) **Bold and italics** denotes elevation presented is the maintenance hole bottom elevation as it was dry at the time of utilized for monitoring.
- 5) For waste cells with two maintenance holes, the maintenance hole farthest into the waste cell was monitored.
- 6) Revised measuring point elevations for Existing Site manholes and leachate wells, as well as updated ground surface elevations reflect the findings of an updated survey (2016).
- 7) PS7 installed in September 2022.
- 8) Existing Ground Surface Elevation and Existing Bottom of Ditch Elevation for MH16, MH17, and MH18 remeasured on November 3, 2022.

Table 3
Leachate Elevation Trends
Twin Creeks Environmenal Centre - 2022 Annual Monitoring Report

Monitor	Long-Term Trend (Includes Historical Data)									
Designation	Constant	Decreasing	Increasing	Fluctuating	Comments					
Cell 3S										
MH3SA				X	Increasing to 2018 and since fluctuating					
MH3SB				X	Decreasing to 2017 and since fluctuating.					
MH3SC				X	Fluctuating since 2017.					
MH3SD				X	Fluctuating since 2017.					
MH3SE	X			X	Increasing to 2017 and since constant to fluctuating.					
MH3SF	X			X	Decreasing to 2017 and since constant to fluctuating.					
Central Fill Area										
OW51A-15	X			X	Constant to fluctuating since 2005.					
MH4A				X	Fluctuating since 2004.					
MH4B				X	Fluctuating since 2015.					
MH5				X	Fluctuating since 2007.					
MH6			Χ	X	Decreasing to 2017 and since increasing to fluctuating.					
MH7		X		X	Fluctuating to decreasing since 2009.					
MH8				X	Fluctuating since 2005.					
МН9				X	Decreasing to 2012 and since fluctuating.					
MH10		X		X	Fluctuating to 2018 and since decreasing.					
MH11				X	Decreasing to 2011 and since fluctuating.					
MH12		X		X	Fluctuating to 2018 and since decreasing.					
SUMP			X	X	Fluctuating to 2016 and since increasing to fluctuating					
South Fill Area										
OW22A-10				X	Fluctuating since 2005.					
OW53-10				X	Increasing to 2014 and since fluctuating.					
MH16				X	Increasing to 2010 and since fluctuating.					
MH17				X	Increasing to 2010 and since fluctuating.					
MH18				X	Increasing to 2010 and since fluctuating.					

## NOTES:

- 1) Elevations can show more than one trend.
- 2) For waste cells with two maintenance holes, the maintenance hole furthest within the waste was utilized for comparison.

Table 4
Groundwater Elevation Trends
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Monitor				ides Historical	Data)
Designation	Constant	Decreasing	Increasing	Fluctuating	Comments
ctive Aquitard					
OW16-6*				X	Seasonal
OW17-4				X	Seasonal
OW40D-4*				X	Seasonal
OW54A-4*			X	X	Seasonal
OW56-4			Х	X	Seasonal
OW57-4			Х	X	Seasonal
OW58-6*			Х	X	Seasonal
OW59-6*				X	Seasonal
OW60-4				X	Seasonal
OW61-4				X	Seasonal
OW62-5*				X	Seasonal
OW67-4				X	
OW68-5				X	Seasonal
OW69-5			Х	Х	
OW70B-5*				X	Seasonal
OW71A-5*				X	Seasonal
OW72-6			Х	X	Seasonal
OW73-6				X	Seasonal
OW75-3				X	Seasonal
OW76-5				X	Seasonal
OW77-4				X	Scasoriai
OW78-4				X	
OW79-5				X	Seasonal
OW80-3				X	Scasorial
OW81-5				X	
OW82-5				Λ	Ins
OW83-5					Ins
OW84-6					
	d Cand				Ins
nterstadial Silt an	ia Sana	l e		V	Caranal
OW16-7				X	Seasonal
OW40A-7*				X	Seasonal
OW46-7			Х	X	
OW47-6				X	
OW54-10		Х		X	Decreasing since 2009
OW57-15				X	
OW58-14				X	Decommissioned 2016
OW58-17				X	Installed April 2014
OW60-8				X	Seasonal
OW61-6				X	
OW62-7				X	
OW67-11				X	Fluctuating since 2018
OW72-10				X	Fluctuating since 2018
OW73-9				X	Fluctuating since 2018
OW75-7				X	-
OW78-6				Х	
OW79-7				X	Seasonal
OW80-6				X	Seasonal
OW81-7				X	

Monitor	Long-Term Trend (Includes Historical Data)										
Designation	Constant	Decreasing	Increasing	Fluctuating	Comments						
OW82-14					Ins						
OW83-9					Ins						
OW84-11					Ins						
Interface Aquifier											
OW17-30				X							
OW19-29				X							
OW39-26				X	Well damaged in 2016						
OW39A-26			X	X	Seasonal						
OW40A-28*		X		X	Decreasing since 2015						
OW49-29		X		X	Decreasing since 2009						
OW60-25		X		X							
OW61-26				X							
OW62-30				X							
OW79-26			X	X	Increasing since 2018						
OW80-27			Х	X	Increasing since 2018						
OW81-27			Х	X	Increasing since 2020						
OW82-28					Ins						
OW83-29					Ins						
OW84-31					Ins						

- 1) Seasonal denotes a seasonal trend with elevated elevations during prolonged periods of precipitation, typically in the spring and summer months, and lower elevations during drier periods, typically the fall and winter months.
- 2) Water elevations can show more than one trend.
- 3) Ins denotes insufficient data. At least three data points are required to establish a trend.
- 4) * denotes trends consider historical data from former monitoring wells at that location.
- 5) Monitoring well OW58-14 was decomissioned prior to the fall 2016 monitoring event.
- 6) Monitoring well OW39-26 was noted to have been damaged during the fall 2016 monitoring event and decomissioned April 2017.
- 7) OW81-5, OW81-7 and OW81-27 installed in June 2019. Insufficient data collected yet to track trend.
- 8) OW82-5, OW82-14, OW82-28, OW83-5, OW83-9, OW83-29, OW84-6, OW84-11, and OW84-31 installed in June 2022.

Table 5
Estimated Leachate Storage Volumes
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

	Approximate	Approximate Original Ground	Groundwater Reference	Estimated Average Base	May 2022 Average				Estin	ated Volu	me in Was	te Cells (Me	asured fro	om Cell Bas	se) (m³)				Change in Volume Within					Volume A	bove Grou	ındwater F	Reference	Level (m³)					Change in Volume Above Groundwater Reference
Fill Area	Area	Surface Elevation	Elevation	of Waste	Leachate	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	the Waste Cells From May	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	MAY	Level From May 2021 to
	(ha)	(m asl)	(m asl)	(m asl)	Elevation	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020	2021	2022	2021 to May 2022 (m ³ )	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020	2021	2022	May 2022 (m³)
West Cell (Sump)	6.3	238.3	237.1	235.5	242.4	103,320	118,440	110,628	136,332	87,948	71,568	148,680	164,052	195,552	242,928	122,724	161,784	174,636	12,852	63,000	78,120	70,308	96,012	47,628	31,248	108,360	123,732	155,232	202,608	82,404	121,464	134,316	12,852
Cell 3 (OW51A-15)	1.5	239.9	239.2	235.2	236.3	4,200	4,980	5,220	5,100	7,800	6,180	4,800	4,920	5,100	6,360	5,640	7,620	6,600	-1,020	0	0	0	0	0	0	0	0	0	0	0	0	0	0
Cell 4 (MH4B)	2.1	240.8	238.5	234.1	239.5	52,080	62,664	29,148	36,120	30,492	37,212	46,200	46,452	40,404	46,368	43,848	44,856	45,024	168	11,760	25,704	0	0	0	252	9,240	9,492	3,444	9,408	6,888	7,896	8,064	168
Cell 5 (MH5A)	2.2	241.2	237.0	236.5	239.7	7,040	0	7,832	28,776	22,264	0	14,520	28,160	0	28,512	27,984	17,600	28,424	10,824	0	0	3,432	24,376	17,864	0	10,120	23,760	0	24,112	23,584	13,200	24,024	10,824
Cell 6 (MH6A)	2.0	240.0	237.3	236.5	240.2	4,000	8,080	5,280	9,680	10,880	1,600	0	0	3,840	28,880	16,720	16,000	29,280	13,280	0	1,680	0	3,280	4,480	0	0	0	0	22,480	10,320	9,600	22,880	13,280
Cell 7 (MH7A)	1.9	240.7	238.8	236.5	237.2	15,960	10,564	19,076	2,736	5,548	17,784	9,576	18,848	12,920	23,484	13,984	10,792	4,940	-5,852	1,520	0	1,596	0	0	304	0	1,368	0	6,004	0	0	0	0
Cell 8 (MH8B)	1.9	240.0	239.3	237.0	239.2	760	10,108	19,380	17,480	3,496	1,672	9,880	18,392	17,252	27,968	29,868	22,572	16,872	-5,700	0	0	1,900	0	0	0	0	912	0	10,488	12,388	5,092	0	-5,092
Cell 9 (MH9A)	1.9	241.2	239.0	237.5	239.2	15,200	17,936	8,816	7,296	7,296	8,968	5,700	13,072	1,064	19,456	14,440	12,692	13,224	532	6,080	6,536	0	0	0	0	0	1,672	0	8,056	3,040	1,292	1,824	532
Cell 10 (MH10)	1.9	241.5	239.3	236.5	236.7	7,600	22,496	21,280	5,320	5,092	8,588	28,500	34,200	28,120	34,276	4,636	4,788	1,368	-3,420	0	1,216	0	0	0	0	7,220	12,920	6,840	12,996	0	0	0	0
Cell 11 (MH11A)	1.9	244.3	240.8	237.8	238.7	20,520	7,068	15,884	14,516	15,352	19,076	23,560	21,584	20,292	25,156	11,324	17,860	7,068	-10,792	9,120	0	0	0	0	0	760	0	0	2,356	0	0	0	0
Cell 12 (MH12A)	0.6	242.5	241.8	236.5	236.7	2,400	7,152	6,768	4,008	1,512	2,640	8,880	10,800	8,856	11,016	1,488	1,488	456	-1,032	0	0	0	0	0	0	0	0	0	0	0	0	0	0
Cell 3S (MH3SA/B/C/D/E/F)	1.1	238.6	238.3	235.2	239.9	7,964	16,060	13,640	16,588	5,412	10,340	14,520	4,752	16,368	21,831	17,431	21,120	20,460	-660	0	2,420	0	2,948	0	0	880	0	2,728	8,191	3,791	7,480	6,820	-660
South Cell (MH16/17/18)	5.4	239.4	238.1	235.0	238.0	54,000	62,640	62,640	67,176	55,296	60,912	49,680	46,224	55,296	70,992	65,304	60,912	65,016	4,104	0	0	0	0	0	0	0	0	0	4,032	0	0	0	0
					Total	295,044	348,188	325,592	351,128	258,388	246,540	364,496	411,456	405,064	587,227	375,391	400,084	413,368	13,284	91,480	115,676	77,236	126,616	69,972	31,804	136,580	173,856	168,244	310,731	142,415	166,024	197,928	31,904

1) Average leachate elevations are from May 2, 2022.

2) Assumes a waste porosity of 40%.

3) NA indicates data not available, owing to date of manhole installation.

4) Revised original ground surface elevations reflect the findings of an updated survey in 2016.

5) Estimated Leachate Storage Volume (m³) for Total in Waste Cell = Average Leachate Elevation (m) - (Estimated Average Base of Waste (ha) * 10,000 (m²)) * 0.4 (assumed porosity).

6) Groundwater Reference Elevation is a fixed value as of 2011 based on historic (2003-2007) data.

7) Volume Above Groundwater Reference Level based on similar reference level for post-2010 volumes. Pre-2010 volumes based on spring average level.

Table 5
Estimated Leachate Storage Volumes
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Fill Area	Approximate Area	Approximate Original Ground Surface Elevation	Groundwater Reference Elevation	Estimated Average Base of Waste	Nov 2022 Average Leachate	( ell Base) (m3)			Change in Volume Within	Change in Volume Within the Waste Cells From Nov				Volume Above Groundwater Reference Level (m3)			
	(ha)	(m asl)	(m asl)	(m asl)	Elevation (m asl)	NOV 2018	NOV 2019	NOV 2020	NOV 2021	NOV 2022	2021 to Nov 2022 (m ³ )	NOV 2018	NOV 2019	NOV 2020	NOV 2021	NOV 2022	Level From Nov 2021 to Nov 2022 (m ³ )
West Cell (Sump)	6.3	238.3	237.1	235.5	242.6	217,224	177,408	139,356	180,432	177,912	-2,520	176,904	137,088	99,036	140,112	137,592	-2,520
Cell 3 (OW51A-15)	1.5	239.9	239.2	235.2	236.6	5,040	6,000	6,840	6,540	8,280	1,740	0	0	0	0	0	0
Cell 4 (MH4B)	2.1	240.8	238.5	234.1	239.4	47,376	40,572	44,772	49,140	44,268	-4,872	10,416	3,612	7,812	12,180	7,308	-4,872
Cell 5 (MH5A)	2.2	241.2	237.0	236.5	236.8	26,136	0	0	30,360	2,552	-27,808	21,736	0	0	25,960	0	-25,960
Cell 6 (MH6A)	2.0	240.0	237.3	236.5	240.2	2,560	30,640	11,360	31,520	29,840	-1,680	0	24,240	4,960	25,120	23,440	-1,680
Cell 7 (MH7A)	1.9	240.7	238.8	236.5	236.6	10,412	3,952	456	4,408	760	-3,648	0	0	0	0	0	0
Cell 8 (MH8B)	1.9	240.0	239.3	237.0	238.9	17,556	6,992	6,308	15,580	14,440	-1,140	76	0	0	0	0	0
Cell 9 (MH9A)	1.9	241.2	239.0	237.5	237.8	4,712	7,752	1,596	9,728	1,900	-7,828	0	0	0	0	0	0
Cell 10 (MH10)	1.9	241.5	239.3	236.5	236.8	33,668	0	0	1,900	2,356	456	12,388	0	0	0	0	0
Cell 11 (MH11A)	1.9	244.3	240.8	237.8	239.8	22,952	7,296	4,484	7,372	15,276	7,904	152	0	0	0	0	0
Cell 12 (MH12A)	0.6	242.5	241.8	236.5	236.8	10,632	0	0	576	792	216	0	0	0	0	0	0
Cell 3S (MH3SA/B/C/D/E/F)	1.1	238.6	238.3	235.2	239.8	19,323	21,685	11,730	19,351	20,108	757	5,683	8,045	0	5,711	6,468	757
South Cell (MH16/17/18)	5.4	239.4	238.1	235.0	238.1	59,760	58,608	58,248	64,800	66,312	1,512	0	0	0	0	0	0
					Total	477,351	360,905	285,150	421,707	384,796	-36,911	227,355	172,985	111,808	209,083	174,808	-34,275

- 1) Average leachate elevations are from November 1, 2022.
- 2) Assumes a waste porosity of 40%.
- 3) NA indicates data not available, owing to date of manhole installation.
- 4) Revised original ground surface elevations reflect the findings of an updated survey in 2016.
- 5) Estimated Leachate Storage Volume (m³) for Total in Waste Cell = Average Leachate Elevation (m) (Estimated Average Base of Waste (ha) * 10,000 (m²)) * 0.4 (assumed porosity).
- 6) Groundwater Reference Elevation is a fixed value as of 2011 based on historic (2003-2007) data.
- 7) Volume Above Groundwater Reference Level based on similar reference level for post-2010 volumes. Pre-2010 volumes based on spring average level.

Table 6
Leachate Chemical Summary
Twin Creeks Environmental Centre – 2022 Annual Monitoring Report

	Le	eachate Concentrations		Вас	kground Concent	rations
Parameter	West Central Fill Area (Existing Site) (2008-2022)	Typical Waste Areas (Existing Site) (2008-2022)	Equalization Tank (Expansion Site) (2010-2022)	Groundwater (1984-2001)	Surface Water (2001-2022) SS10	Surface Water (2008-2022) SS16
pH (pH units)	7.3 - 8.0	7.4 - 8.2	7.0 - 8.1	7.0 - 8.3	6.7 - 8.7	7.4 - 8.4
Conductivity (µS/cm)	880 - 16,400	501 – 31,100	2,000 - 26,000	443 – 2,550	62 - 1200	109 - 620
Alkalinity	290 - 7,060	160 - 18,300	1,820 - 10,000	100 - 581	22 - 203	45 - 228
Calcium	76 - 150	19 - 380	79 - 1,400	19 - 250	5.4 - 170	15 - 1300
Magnesium	22 - 390	19 - 450	130 - 530	9 - 261	1.6 - 33	3.3 - 85
Sodium	57 - 2,100	19 - 6,300	270 - 2,700	48 - 199	0.49 - 20	1.4 - 18
Potassium	12 - 620	6.7 - 1,100	81 - 900	1 - 9	1 - 17	3.1 - 60
Chloride	68 - 2,600	11 - 1,000	330 - 3,400	1 - 15	1 - 46	1.2 - 20
Sulphate	0.5 - 81	10 - 410	10 - 220	100 - 1,330	0.5 - 220	<1 - 220
Iron	2.2 - 54	<0.05 - 33	1.6 - 120	<0.1 - 3.3	0.25 - 79	0.3 - 540
DOC	25 - 462	8.3 - 1,480	170 - 4,500	0.7 - 9.8	<0.02 - 0.26	
Ammonia (total)	11.1 - 857	8 - 3,540	57 - 2,000	<0.05 - 7.1	<0.02 - 18.4	0.05 - 0.3
TKN	11 - 930	11 - 3,500	26 - 2,700	<0.1 – 10.9	0.41 - 33	<0.7 - 7.2
Nitrate	<0.1 - <2.0	<0.1 - 6.7	<0.1 - <5.0	<0.1 - 2.7	<0.01 - 102	<0.01 - 1.5
Boron	1.3 - 70	0.67 - 560	1.3 - 28	0.09 - 0.99	0.02 - 0.48	<0.02 - 0.4
Benzene (µg/L)	<0.2 - 361	<0.1 - 12.0	<1 - <10	<1.3		
Toluene (µg/L)	<0.2 - 782	<0.2 - 550	4.2 - 2400	<1.5		
Ethylbenzene (µg/L)	<0.2 - 318	<0.1 - 891	<0.5 - 21	<1.6		
m/p - xylenes (µg/L)	<0.2 - 1,990	<0.1 - 200	2.5 - 55	<3.4		
o - xylene (µg/L)	<0.2 - 1140	<0.1 - 97.4	<0.5 - 22	<2.7		

¹⁾ Background concentrations for groundwater are established from 1984 to 2001 for monitoring wells 250 metres or greater to the west of the Existing Site: OW1-5, OW5-6, OW38-6, OW38-10, OW39-6, OW39-12, OW42-6, OW42-9.

²⁾ Background concentrations for surface water (SS10 and SS16) are established for 2001-2022 data, where available.

³⁾ All data are mg/L unless otherwise specified.

⁴⁾ Blank denotes parameter not tested.

Table 7
Indicator Parameter Concentration Trend Summary - Groundwater
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

	Long-To	erm Trend (Includes Histor	ric Data)
Monitor Designation	Chloride	Nitrate	Boron
	Activ	e Aquitard	
OW16-6*	С	С	С
OW17-4	F	С	С
OW40D-4	С	С	D/F
OW54A-4*	I/F	С	С
OW56-4	С	F	F
OW57-4	С	С	С
OW58-6	С	С	F
OW59-6	С	С	F
OW60-4	С	С	С
OW67-4	F	F	D/F
OW68-5	С	С	С
OW69-5	С	С	F
OW70B-5*	С	С	С
OW71A-5*		С	С
OW72-6	C	C	C
OW73-6	C	C	F
OW79-5	F	С	С
OW80-3	F	C	C
OW81-5	C	C	C
OW82-5	ID	ID	ID
OW83-5	ID	ID	ID
OW84-6	ID	ID	ID
011010		al Silt and Sand	
OW16-7	C	C	С
OW40A-7	C	C	C
OW46-7	C	C	F/I
OW47-6	C	F	F
OW54-10	C	C	C
OW57-15	C	C	C
OW58-17*	C	C	C
OW60-8	C	C	F
OW67-11	F	F	F
OW72-10	C	C	C
OW73-9	C	С	F
OW79-7	F	C	C
OW80-6	' F	С	С
OW81-7	<u>·</u>	С	С
OW82-14	ID	ID	ID
OW83-9	ID ID	ID	ID
OW84-11	ID	ID	ID
J07 11		ace Aquifer	ID.
OW19-29	С	C	F
OW19-29 OW39A-26*	F	С	F
OW49-29	C	C	F
OW79-26	C	F	C
OW80-27	C	C	F
OW81-27		С	C
OW81-27 OW82-28	ID	ID	ID
OW83-29	ID	ID	ID
OW83-29 OW84-31	ID ID	ID	ID
Cemetery Well	C	F F	C
Cemetery well	<u> </u>	F	

- 1) C denotes constant trend.
- 2) D denotes decreasing trend.
- 3) I denotes increasing trend.
- 4) F denotes fluctuating trend.
- 5) S denotes seasonal pattern.
- 6) ID denotes insufficient data for interpretation.
- 7) Data are presented in Table H-4 Appendix H .
- 8) Historic data trends assume nitrate+nitrite represents nitrate.
- 9) Concentrations can show more than one trend.
- 10)  $\,^{\star}$  denotes considers historic data from decommissioned monitoring well.

Table 8
Groundwater Trigger Concentration / Comparison Summary - Active Aquitard
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Monitor				
Designation		Chloride	Nitrate	Boron
	Active	Aquitard		
	Trigger Concentration (mg/L)	106	2.3	1.1
01446.6	May 2022	48	<0.10	0.16
OW16-6	November 2022	45	0.15	0.18
OW17-4	May 2022	32	0.16	0.24
OW17-4	November 2022	28	0.14	0.33
OW40D-4	May 2022	6.2	<0.10	0.33
OME 4A 4	May 2022	28	<0.10	0.18
OW54A-4	November 2022	25	<0.10	0.24
OWEG A	May 2022	5.8	0.36	0.35
OW56-4	November 2022	7.4	0.14	0.50
014/57.4	May 2022	6.0	<0.10	0.35
OW57-4	November 2022	6.1	<0.10	0.45
OWER C	May 2022	5.5	0.47	0.41
OW58-6	November 2022	4.6	<0.10	0.69
OW59-6	May 2022	6.2	<0.10	0.72
Ovv59-6	November 2022	6.4	<0.10	0.74
OW60-4	May 2022	8.3	<0.10	0.04
OW67-4	May 2022	23	0.20	0.68
OVV07-4	November 2022	60	0.16	0.21
OW68-5	May 2022	9.9	<0.10	0.08
Ovv68-5	November 2022	9.8	<0.10	0.10
OW69-5	May 2022	12	<0.10	0.94
Ovv69-5	November 2022	11	<0.10	1.1
OW70B-5	May 2022	6.8	<0.10	0.42
OW/UB-5	November 2022	7.2	<0.10	0.51
OW71A-5	May 2022	30	<0.10	0.12
OW/IA-5	November 2022	32	<0.10	0.13
OW72-6	May 2022	6.5	<0.10	0.43
UW/2-0	November 2022	4.8	<0.10	0.70
OW73-6	May 2022	12	<0.10	1.0
UW/3-6	November 2022	8.2	<0.10	0.86
OW79-5	May 2022	26	<0.10	0.06
OW80-3	May 2022	67	0.22	0.03
OW81-5	May 2022	20	<0.10	0.54

## NOTES:

- 1) Trigger Concentrations based on Table 4 of the 2007 Landfill EMP.
- 2) Bolded text and shading denotes concentration exceeds trigger concentration.
- 3) Italics denotes that per MECP approval the parameter's concentration is not evaluated against the trigger concertation.
- 4) Monitoring well nest OW81 installed in June 2019.
- 5) Monitoring well nests OW82, OW83, and OW84 were iinstalled in June 2022 and will be monitored in May 2023.

Table 9
Groundwater Trigger Concentration Comparison Summary - Interstadial Silt and Sand
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Monitor De	esignation	Chloride	Nitrate	Boron
	Interstadial Silt a	and Sand		
	Trigger Concentration (mg/L)	116	2.3	2.1
OW16-7	May 2022	7.4	0.57	0.25
OW 16-7	November 2022	6.6	0.29	0.27
OW40A-7	May 2022	7.2	0.40	0.61
OW46-7	May 2022	14	0.15	0.58
OVV46-7	November 2022	13	0.17	0.80
OW47-6	May 2022	14	0.15	0.58
OW47-6	November 2022	5.9	0.33	1.0
OW54-10	May 2022	9.2	0.57	0.95
OVV54-10	November 2022	8.8	0.27	1.0
OW57-15	May 2022	8.8	0.80	1.2
OVV37-13	November 2022	8.7	<0.10	1.2
OW58-17	May 2022	8.4	0.25	1.4
OVV36-17	November 2022	9.5	<0.10	1.3
OW60-8	May 2022	8.8	<0.10	0.85
OW67-11	May 2022	30	0.26	0.47
OW67-11	November 2022	25	0.10	0.58
OW72-10	May 2022	5.7	0.38	1.2
OW/2-10	November 2022	5.4	<0.10	1.1
OW73-9	May 2022	8.3	0.61	1.2
UW/3-9	November 2022	7.7	0.60	1.3
OW79-7	May 2022	160	<0.10	0.21
OW80-6	May 2022	4.2	<0.10	0.05
OW81-7	May 2022	200	0.16	0.54

- 1) Trigger Concentrations based on Table 4 of the 2007 Landfill EMP.
- 2) Bolded text and shading denotes concentration exceeds trigger concentration.
- 3) Italics denotes that per MECP approval the parameter's concentration is not evaluated against the trigger concertation.
- 4) Chloride trigger concentration of 116mg/L is not applicable to OW79-7 and OW80-6
- 5) Monitoring well nest OW81 installed in June 2019.
- $6) \, Monitoring \, well \, nests \, OW82, \, OW83, \, and \, OW84 \, were \, iinstalled \, in \, June \, 2022 \, and \, will \, be \, monitored \, in \, May \, 2023.$

Table 10
Groundwater Trigger Concentration Comparison Summary - Interface Aquifer
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Monitor Designation		Chloride	Nitrate	Boron
	Interface Aquifer			
	Trigger Concentration (mg/L)	134	2.3	2.6
OW19-29	May 2022	29	<0.10	2.3
OW39A-26	May 2022	42	<0.10	1.8
OW49-29	May 2022	27	<0.10	1.7
OW79-26	May 2022	28	0.16	1.6
OW80-27	May 2022	47	<0.10	1.9
OW81-27	May 2022	30	<0.10	1.9
Cemetery Well	May 2022	3.8	0.33	0.03

- 1) Trigger Concentrations based on Table 4 of the 2007 Landfill EMP.
- 2) Bolded text and shading denotes concentration exceeds trigger concentration.
- 3) Monitoring well nest OW81 installed in June 2019.
- 4) Monitoring well nests OW82, OW83, and OW84 were iinstalled in June 2022 and will be monitored in May 2023.

Table 11
Indicator Parameter Concentration Trend Summary - Surface Water
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Monitoring Station	Long	-Term Trend (In	cludes Historic	Data)
Wollitoring Station	Chloride	Ammonia	Boron	Zinc
SS1	F	С	F	F
SS10	F	С	F	F
SS16	F	С	F	С
SP1	F	F	F	С
SP2	F	С	F	С
SP3	F	F	С	С
SP4	F	С	С	С

- 1) C denotes constant trend.
- 2) D denotes decreasing trend.
- 3) I denotes increasing trend.
- 4) F denotes fluctuating trend.
- 5) S denotes seasonal pattern.
- 6) Data are presented in Table I-2, Appendix I.
- 7) INS denotes there is insufficient data available for interpretation.
- 8) NC denotes sampling station not yet constructed.

Table 12
Surface Water Trigger Concentrations
Twin Creeks Environmental Centre - 2022 Annual Monitoring Report

Parameter	Units	PWQO	Background Station SS10 and SS16 90% Trigger Concentration	Number of Samples for Background Station SS10	Number of Samples for Background Station SS16	Post 2021 Trigger Concentration	Post 2022 Trigger Concentration
			Trigger Conce	entrations/Levels for Complia	ince Points		
Chloride	mg/L	210*	20.0	64	47	210	210
Ammonia (unionized)	mg/L	0.020	0.010	64	47	0.020	0.020
Phenols	mg/L	0.001	0.004	64	47	0.004	0.004
Boron	mg/L	0.20	0.17	64	47	0.20**	0.20**
Nickel	mg/L	0.025	0.027	64	47	0.027	0.027
Chromium (total)	mg/L	0.0089	0.024	64	47	0.024	0.024
Zinc	mg/L	0.02	0.06	64	47	0.06	0.06

- 1) PWQO denotes Provincial Water Quality Objectives (1994) with updates.
- 2) * denotes recommended concentration by Environment Canada.
- 3) ** It is noted that the boron trigger concentration for SP1 is 0.39 mg/L, per the MOECC letter dated May 18, 2012.
- 4) Trigger concentrations are calculated annually from SS10 and SS16, per MOECC letter dated February 27, 2014.

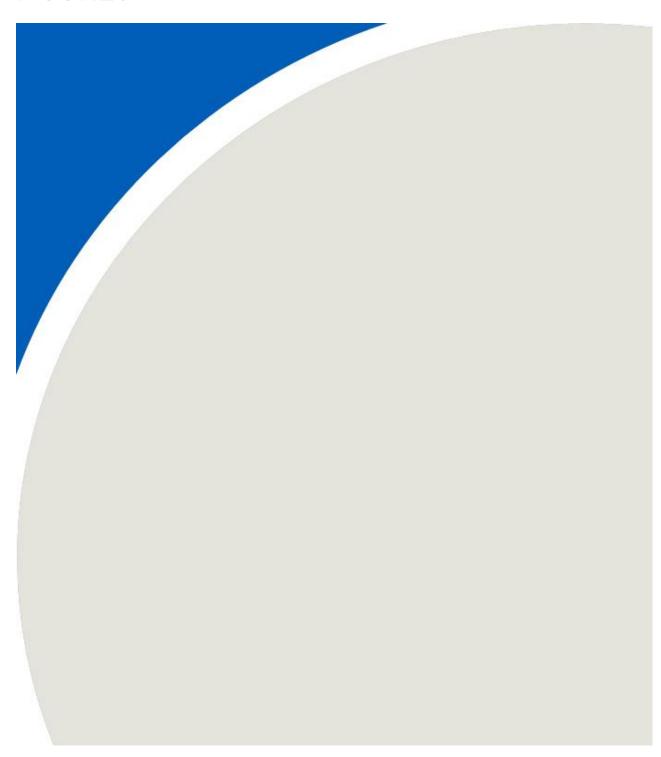
Monitoring Locations	Parameters	Frequency
Leachate		
PS1, PS3, PS5, PS7, MH3S, MH4, MH5, MH6, MH7, MH8, MH9, MH10, MH11, MH12, MH16, MH17, MH18, OW22A-10,	Leachate Levels	May and November
OW51A-15, OW53-10, Sump, LW1, LW2, LW3, LW4, LW5, LW6 PS1, PS3, PS5, PS7	Leachate Levels	Daily
PS1, PS3, PS5, PS7, South Fill Area (MH18), West Central Fill Area (Sump), Central Fill Area (Composite of MH3, MH4, MH5,		July
MH6, MH7, MH9, MH11)	SW, LS	May
	BOD ₅ , DOC, phosphorus (total),	Out to the
Equalization Tank	TKN, BTEX, pH	Quarterly
Treated Leachate Effluent	PLIL-SW, SLIL-SW, LS	May and November
Treated Leachate Emident	Discharge Rates, COD, pH,	
	turbidity	Daily
Treatment Plant Effluent	Chloride, CBOD ₅ , BOD ₅ , DOC,	Weekly
Treatment Plant Emberit	BTEX, ammonia, pH PLIL-GW, SLIL-GW, PLIL-SW, SLIL-	
	SW, LS	Monthly
Treated Lagrage Tampayan, Stayana Calla	PCB, organochlorines	May and November
Treated Leachate Temporary Storage Cells :  Cells 1 and 2	Discharge Rates	Daily
Cell 1 Inlet, Cells 1 and 2	Chloride, CBOD ₅ , BOD ₅ , DOC,	Weekly
Cells 1 and 2	BTEX, ammonia, pH DO, pH, alkalinity, DOC	Weekly
	PLIL-GW, SLIL-GW, PLIL-SW, SLIL-	
Cells 1 and 2	SW, LS	Monthly
Cell 1	Biomonitoring	May and November
PS2, PS4, PS6, PS8	Groundwater Levels	Monthly
Active Aquitard		
OW16-6, OW17-4, OW40D-4, OW54A-4, OW56-4, OW57-4, OW58-6, OW59-6, OW60-4, <b>OW61-4, OW62-5</b> , OW67-4, OW68-		
5, OW69-5, OW70B-5, OW71A-5 [†] , OW72-6, OW73-6, <b>OW75-3, OW76-5, OW77-4, OW78-4, OW79-5</b> , OW80-3, OW81-5,	Groundwater Levels	May and November
OW82-5, OW83-5, OW84-6, <b>OW85-5</b> , P1, P2, P3		
OWIG C ONIG A OWE A A OWE C A OWE AND A OWE A OWE AND A OWE		
OW16-6, OW17-4, OW54A-4, OW56-4, OW57-4, OW58-6, OW59-6, <b>OW61-4, OW62-5</b> , OW67-4, OW68-5, OW70B-5, OW71A-5*, OW72-6, OW73-6, <b>OW75-3, OW76-5, OW77-4, OW78-4</b>	PLIL-GW, SLIL-GW	May and November
OW40D-4, OW60-4, OW79-5, OW80-3, OW81-5, OW82-5, OW83-5, OW84-6, <b>OW85-5</b>	PLIL-GW, SLIL-GW	May
OW16-6, <b>OW61-4, OW62-5, OW75-3, OW78-4</b>	Volatiles	May and November
OW17-4, OW40D-4, OW54A-4, OW56-4, OW57-4, OW58-6, OW59-6, OW60-4, OW67-4, OW68-5, OW69-5, OW70B-5, OW71A-5*, OW72-6, OW73-6, <b>OW76-5, OW77-4</b> , OW79-5, OW80-3, OW81-5, OW82-5, OW83-5, OW84-6, <b>OW85-5</b>	Volatiles	May
0W/1A-5^, 0W/2-6, 0W/3-6, <b>0W/6-3, 0W/7-4</b> , 0W/9-5, 0W80-3, 0W81-5, 0W82-5, 0W83-5, 0W84-6, <b>0W85-5</b>		
Interstadial Silt and Sand		
OW16-7, OW40A-7, OW46-7, OW47-6, OW54-10, OW57-15, OW58-17, OW60-8, <b>OW61-6, OW62-7</b> , OW67-11, OW72-10,		
OW73-9, <b>OW75-7, OW78-6</b> , OW79-7, OW80-6, OW81-7, OW82-14, OW83-9, OW84-11, <b>OW85-8</b>	Groundwater Levels	May and November
OW46-7, OW47-6, OW54-10, OW57-15, OW58-17, OW67-11, OW72-10, OW73-9	PLIL-GW, SLIL-GW	May and November
OW16-7, <b>OW61-6, OW62-7, OW75-7, OW78-6</b>	PLIL-GW, SLIL-GW, volatiles	May and November
OW40A-7 OW60-8, OW79-7, OW80-6, OW81-7, OW82-14, OW83-9, OW84-11, <b>OW85-8</b>	PLIL-GW, SLIL-GW	May
OW40A-7, OW46-7, OW47-6, OW54-10, OW57-15, OW58-17, OW60-8, OW67-11, OW72-10, OW73-9, OW79-7, OW80-6,	V 1	
OW81-7, OW82-14, OW83-9, OW84-11, <b>OW85-8</b>	Volatiles	May
Interface Aquifer		
OW17-30, OW19-29, OW39A-26, OW40A-28, OW49-29, OW60-25, <b>OW61-26</b> , <b>OW62-30</b> , OW79-26, OW80-27, OW81-27,	Groundwater Levels	May and November
OW82-28, OW83-29, OW84-31		
OW19-29, OW39A-26, OW49-29, OW79-26, OW80-27, OW81-27, OW82-28, OW83-29, OW84-31, Cemetery Well	PLIL-GW, SLIL-GW	May
OW19-29, OW39A-26, OW49-29, OW79-26, OW80-27, OW81-27, OW82-28, OW83-29, OW84-31, Cemetery Well	Volatiles	Biennial - May 2024
Background Station	Flow Rates	
		Quarterly after 10 mm precipitation events.
SS10, SS16	PLIL-SW, SLIL-SW, nitrite	Greater than 1 month intervals between sampling.
	LS-SW	Spring Precipitation Event
	Biomonitoring	Spring Precipitation Event
Sedimentation Ponds (Discharge Points)		
	Flow Rates	Quarterly after 10 mm precipitation events.
	PLIL-SW, SLIL-SW, nitrite	Greater than 1 month intervals between sampling.
SP1, SP2, SP3, SP4		· -
۳ ال رک ال ۱٫۵ اد را الد	LS-SW, volatiles, semi-volatiles	Quarterly after 10 mm precipitation events.
	Lo-344, volatiles, Settii-Volatiles	Greater than 1 month intervals between sampling.
	Biomonitoring	Spring Precipitation Event
Western Site Boundary Complian	ce Point	
	Flow Rates	Quarterly after 10 mm precipitation events.
	PLIL-SW, SLIL-SW, nitrite	Greater than 1 month intervals between sampling.
SS1		
	LS-SW, volatiles, semi-volatiles	Quarterly after 10 mm precipitation events.
		Greater than 1 month intervals between sampling.
Poplar Tree Plantation Land Applica	Biomonitoring	Spring Precipitation Event
Popial Tree Plantation Land Applica	Flow Rates	
		Quarterly after 10 mm precipitation events.
	PLIL-SW, SLIL-SW, nitrite	Greater than 1 month intervals between sampling.
SS17A, SS17B, SS18A, SS18B		Quarterly after 10 mm precipitation events.
	LS-SW, volatiles, semi-volatiles	Greater than 1 month intervals between sampling.
	Biomonitoring	Spring Precipitation Event
Compost Facility (if construct		Spring i recipitation Event
2010	PLIL-SW, SLIL-SW, nitrite, BOD ₅ ,	a
SS19	TSS, Total Coliform, Fecal	Prior to water use
	Coliform, E. Coli	

Monitoring Locations	Parameters	Frequency	
Landfill Gas Monitoring			
Landfill Cap	Inspections	Monthly (April to November)	
GP1A, GP2, GP3, GP4, GP5, GP6, GP7, GP8, GP9, GP10	Methane Gas	January, February, March, April, July, November, December	

- 1) PLIL-GW indicates: chloride, nitrate, boron.
- 2) SLIL-GW indicates: alkalinity, sulphate, calcium, magnesium, potassium, sodium, barium, cadmium, iron, lead, DOC, TDS, ammonia (total), TKN, pH, conductivity. Field parameters of pH, conductivity, temperature, turbidity.
- 3) PLIL-SW indicates: chloride, ammonia (total and unionized), phenols, boron, nickel, chromium (total), zinc.
- 4) SLIL-SW indicates: alkalinity, sulphate, calcium, magnesium, potassium, sodium, total phosphorus, iron, nitrate, TKN, TDS, pH, conductivity. Field parameters of temperature, pH, conductivity, turbidity, DO.
- $5) \ LS \ indicates: \ arsenic, \ barium, \ cadmium, \ copper, \ lead, \ manganese, \ mercury, \ nitrite, \ TSS, \ volatiles, \ semi-volatiles, \ BOD_5, \ COD.$
- 6) LS-SW indicates: arsenic, barium, cadmium, copper, lead, mercury, nitrite, TSS, BOD₅, COD.
- 7) Volatiles should include the following at a minimum: benzene, 1,4-dichlorobenzene, dichloromethane, toluene, ethylbenzene, xylenes, and vinyl chloride.
- 8) Semi-volatiles should include the following at a minimum: 1,2,4-trichlorobenzene, 1,2-dichlorobenzene, 1,3-dichlorobenzene, 1,4-dichlorobenzene, hexachlorobenzene, diethylphthalate, dimethylphthalate, di-n-butyl phthalate, phenol, benzo(a)pyrene, 2,4,6-trichlorophenol, 2,4-dichlorophenol, pentachlorophenol.
- 9) Organochlorines include herbicide and pesticide scan.
- 10) Biomonitoring indicates toxicity testing for Rainbow Trout and Daphnia Magna.
- 11) Biennial indicates every second year.
- 12) QA/QC includes one (1) blind duplicate for each 15 samples or once per event, whichever is greater.
- 13) Surface water samples shall be collected in a downstream to upstream sequence.
- 14) OW84(new) denotes monitoring wells to be installed per EMP dated December 20, 2007.
- 15) Spring denotes: April, May, and June.
- 16) indicates that OW71A-5 is not required as part of the monitoring program, however, obtained data is interpolated for the monitoring well OW67-4, which used to show dry conditions.
- 17) Since the Poplar Plantation is not required to be utilized until a few months prior to the initiation of the treatment plant as operational, monitoring per the EMP and ECA, as well as the Waste and Sewage ECA's that is completed to evaluate the vigour of the Poplar Plantation, is not required. It is recognized that once the Poplar Plantation would be reinitiated.
- 18) Monitoring stations that are currently idle until 2 months prior to the leachate treatment plant being operational, include the following: 1) Surface water stations **SS17A**, **SS17B**, **SS18A**; and 2) Groundwater monitoring locations **OW61**, **OW62**, **OW75**, **OW76**, **OW77**, **OW78**, **AND OW85**.



# **FIGURES**



L/2202861/JdDelwerables/41WasteManagement/Phase 1000/Q4/Q4 Annual Report/03_Figures/F110 F3 Working Files - do not include/20221116 RWDI 2202861 HG Waste TCEC/2022

MAP SOURCE:
AERIAL PHOTO PROVIDED BY GOOGLE EARTH (2011), ACCESSED ON JANUARY 2014.
FIGURE PROVIDED BY GENIVAR INC., FIGURE 2, FILE REF NO. 111-53111-00-F2-SP 12CM Q4, DATED FEBRUARY 2013.

FIGURE 2

SITE PLAN
2022 COMPLIANCE MONITORING PROGRAM

DATE: SCALE

⋚렃

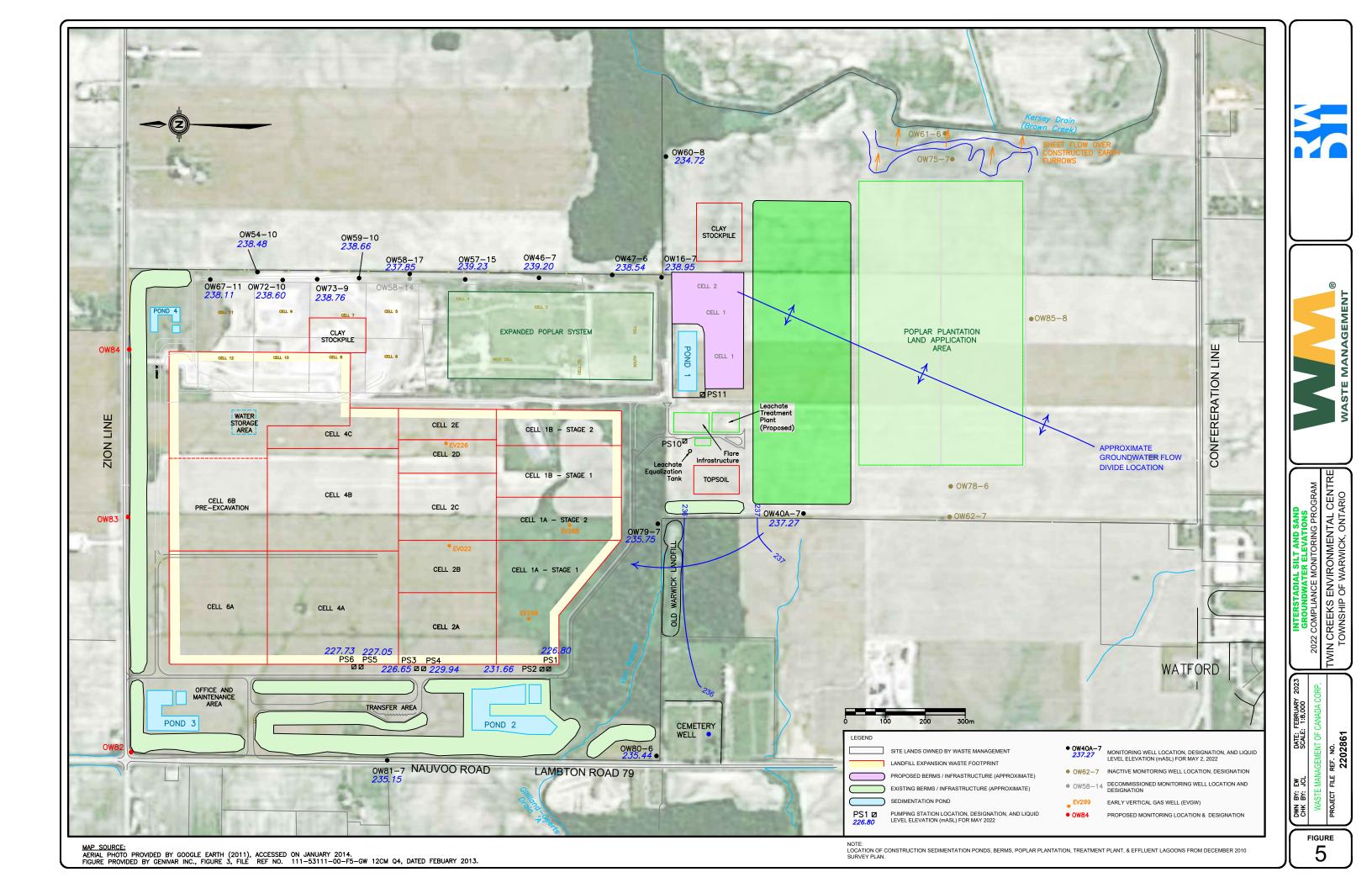
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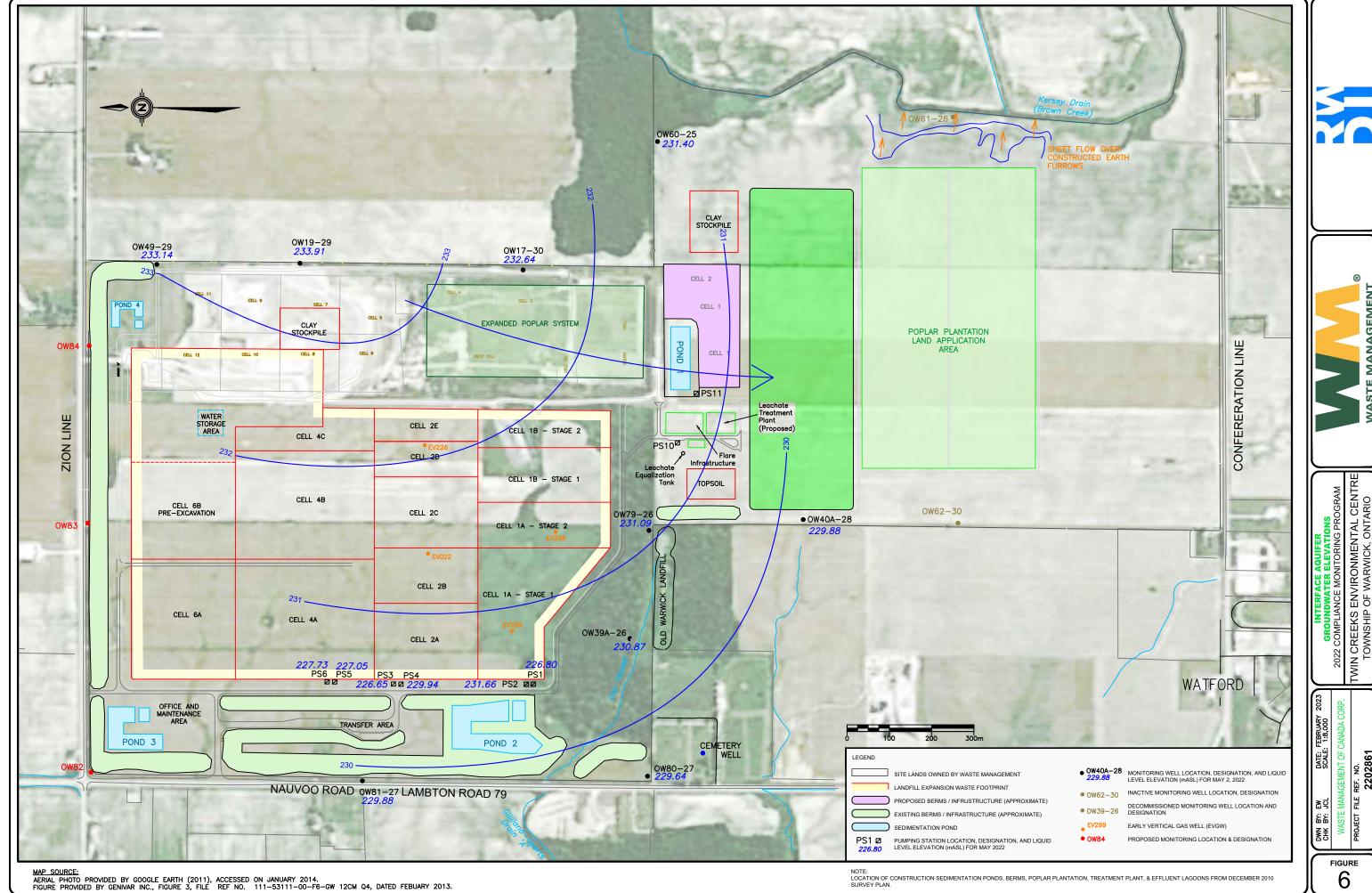
NOTE: LOCATION OF CONSTRUCTION SEDIMENTATION PONDS, BERMS, POPLAR PLANTATION, TREATMENT PLANT, & EFFLUENT LAGOONS FROM NOVEMBER 2010 SURVEY PLAN.

SURFACE WATER SYSTEM
2022 COMPLANCE MONITORING PROGRAM
TWIN CREEKS ENVIRONMENTAL CENTRE
TOWNSHIP OF WARWICK, ONTARIO

DATE: NOVEMBER SCALE: 1:8,000

FIGURE 3





FIGURE

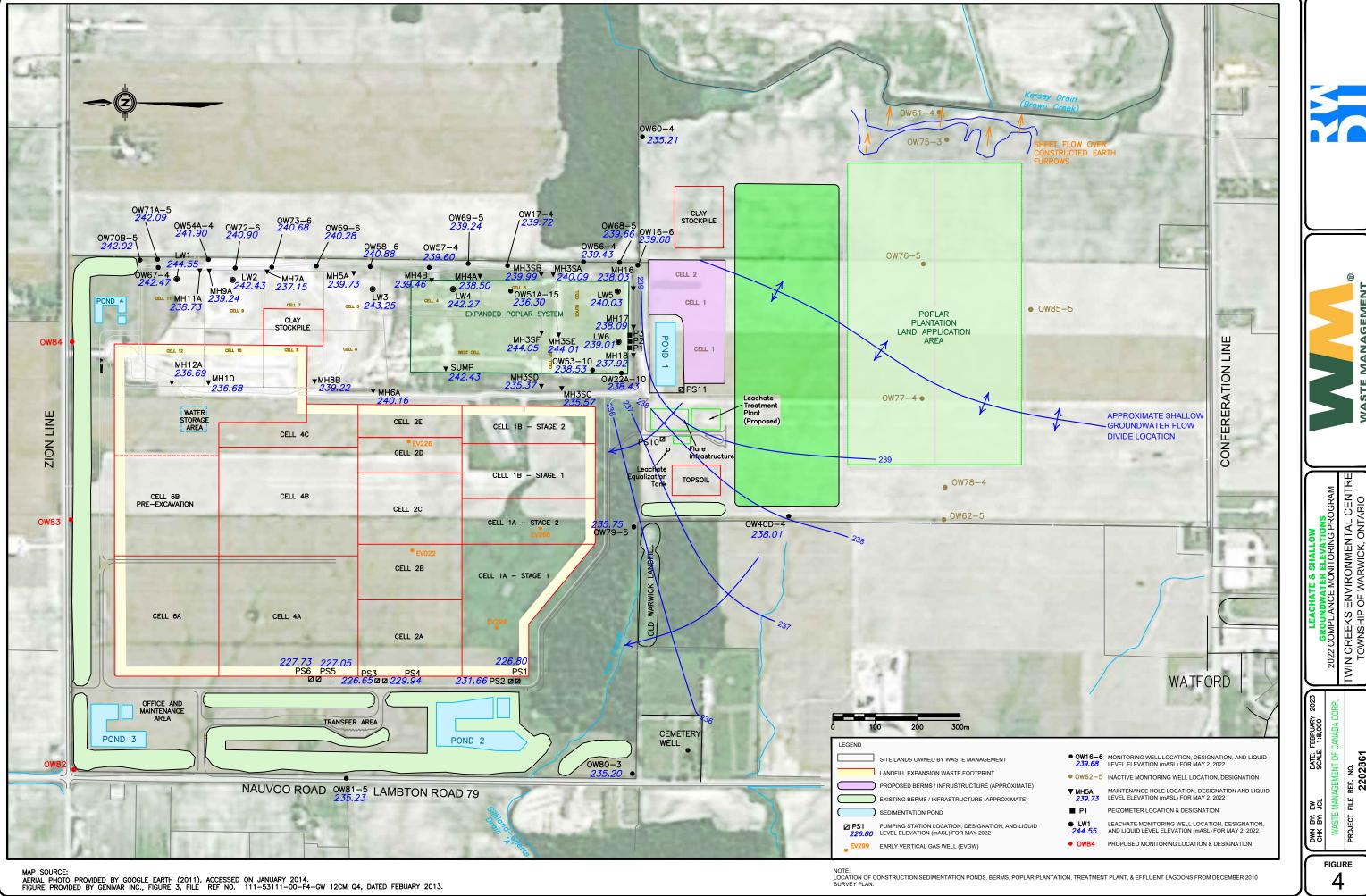


FIGURE 4